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Urinary pesticide concentrations in French adults with low and high organic food consumption: results from the general population-based NutriNet-Santé

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Running title: pesticide exposure and organic consumption

1 Abstract

2 An organic diet may reduce dietary exposure to pesticides but findings based on observational data are 3 scant. We aimed to compare urinary pesticide concentrations between "organic" and "conventional" 4 consumers from the NutriNet-Santé study. Organic food consumption was determined using a self-5 reported food frequency questionnaire. Individuals with a proportion of organic food in the whole diet (in g/d) below 10% were defined as low organic food consumers and those whose proportion was 6 7 above 50% as high organic food consumers. A propensity score matching procedure was then used to 8 obtain two similar subsets of 150 participants, differing mostly by the organic valence of their diet. 9 Urinary pesticide and metabolite concentrations (organophosphorus, pyrethroid and azole compounds) 10 were determined by UPLC-MS/MS, standardized with respect to creatinine. The molar sums of total 11 diethylphosphates, dimethylphosphates and dialkylphosphates were also computed. Differences in 12 distributions across groups were tested using Wilcoxon signed-rank test for matched data. Mean age was 58.5y and 70% of participants were women. Significantly lower urinary levels of 13 14 diethylthiophosphate, dimethylthiophosphate, dialkylphosphates and free 3-phenoxybenzoic acid were 15 observed among organic consumers compared to conventional consumers. Our findings confirm that 16 exposure to certain organophosphate and pyrethroïd pesticides in adults may be lowered by switching 17 from conventional to organic foods. This is particularly of high interest among conventional fruit and 18 vegetables consumers, as their exposure may be the highest.

19 Keywords: dietary exposure; epidemiology; pesticides

20 Introduction

21 Pesticides are widely used to protect crops against harmful organisms and diseases but are also utilized 22 as biocides for non-agricultural purposes. Despite these beneficial aspects, there has been an 23 increasing concern during the last decade that these compounds represent a risk to the general population through residues in food commodities. The toxicological outcomes that have been 24 associated with pesticide exposure include neurological, respiratory, dermatological, digestive, 25 carcinogenic, reproductive and developmental effects, as recently reviewed by several authors ¹⁻³. Five 26 27 pesticides (malathion, glyphosate, parathion, diazinon and tetrachlorvinphos) were classified as probably carcinogenic to humans by the International Agency for Research on Cancer⁴. In addition, a 28 high number of agrochemicals have been documented to affect the endocrine system ^{5,6}, causing 29 reproductive and developmental adverse effects, but also resulting in metabolic disorders ^{7,8}. 30 31 Knowledge on the consequences of exposure levels observed in the general population remains scarce and difficult to interpret ⁹. France is one of the largest users in tons of agricultural pesticides in the 32 European Union¹⁰. Routes of exposure to pesticides are multiple (oral, dermal and respiratory) but 33 34 diet is the main source of pesticide exposure in the general population ¹¹. It is therefore essential to 35 estimate how and to what extent, different dietary consumption patterns - from more or less 36 contaminated food sources (e.g. organically vs. conventionally grown products) may affect exposure.

Organic production and labelling of organic products are held under legal framework of the Council 37 Regulation (EC) No 834/2007, which limits the use of pesticides to a small number (i.e. 35) of natural 38 39 substances while 488 active substances are approved by (EC) No 1107/2009 as pesticides in conventional agriculture in the European Union¹². This exclusion of synthetic pesticides results in a 40 41 significantly lower frequency of (or no) contamination in organic foods when compared to conventional foods, as consistently described in food residue analyses ¹³⁻¹⁶. Some experimental 42 studies, using mostly cross-over design, have been carried out among children ¹⁷⁻¹⁹ and adults ^{20,21}, and 43 all report that the adoption of a diet mainly based on organic foods leads to a significant reduction in 44 pesticide levels in urine (including organophophate pesticides and herbicide 2,4-D¹⁷⁻²¹). Furthermore, 45 the considerable recent growth of organic food market ²² is largely due to consumers' concerns for 46 47 food safety. Indeed, organic products are perceived healthier by consumers than conventional ones

48 mainly because of their absence of pesticide residues ^{23–27}. Yet the extent to which day-life high 49 organic food consumption is related to reduced urinary pesticide concentrations, more specifically in 50 European adults, is not well documented.

51 In this context, the objective of the current study was to test for differences in pesticide exposure, 52 reflected by urinary biomarkers, among adults with low and high self-reported organic food 53 consumption.

54

55 Methods

56 *Study population*

57 The NutriNet-Santé study is an ongoing web-based observational prospective study launched in 58 France in May 2009 on a large sample of adult volunteers. Its general aim is to investigate the 59 relationships between dietary patterns, nutrition and health issues ²⁸. Participants over age of 18y are 60 recruited among the general population by means of vast multimedia campaigns. All questionnaires 61 are completed online using a dedicated website.

The NutriNet-Santé study is conducted according to the Declaration of Helsinki guidelines and was approved by the Institutional Review Board of the French Institute for Health and Medical Research (IRB Inserm n°0000388FWA00005831) and the "Commission Nationale de l'Informatique et des Libertés" (CNIL n°908450/n°909216). Clinicaltrials.gov number: NCT03335644.

On a voluntary basis, participants were also invited to attend health centers for biological sampling and clinical examination (2011-2013). During the visit, volunteers underwent blood and urine sampling as well as a clinical examination including anthropometric measurements. Overall, samples of serum, plasma, buffy-coats and urine were set up for about 20,000 participants of the cohort ²⁹. Electronic and paper written informed consents were obtained from all subjects attending the visit. All procedures were approved by the Consultation Committee for the Protection of Participants in Biomedical Research" (C09-42 on May 5th 2010) and the CNIL (n°1460707).

73

74 Sociodemographic, anthropometric and biological data

To be included in the cohort, participants have to fill in a set of questionnaires providing information on sociodemographic (age, sex, educational level, employment status, place of residence) and lifestyle (smoking status, physical activity) characteristics and health data (menopausal status for women, medical history and medication). Every year thereafter, as part of their follow-up, they are also invited to complete this same set of questionnaires to update their information.

During the clinical visits, systolic and diastolic blood pressure, weight and height were measured using
 standardized procedures ³⁰. Fasting blood glucose, total serum cholesterol, HDL-cholesterol and serum
 triglycerides were routinely measured as previously described ²⁹.

83

84 Assessment of total and organic food consumption

85 Total and organic food consumptions were assessed using an organic semi-quantitative food frequency 86 questionnaire (Org-FFQ). In June 2014, NutriNet-Santé participants were invited to complete the Org-FFQ through the dedicated secured website. The development of the Org-FFQ has been fully 87 described elsewhere ³¹. Briefly, the Org-FFQ consisted of 264 food and beverage items. It was based 88 on a validated FFQ ³², completed by specific questions about organic food consumption frequency. 89 90 Briefly, for each of the 264 items subjects were asked to report their frequency of consumption and the 91 quantity consumed over the past year. Additionally, a five-point Likert-type scale ranging from never 92 to always was used to estimate the frequency of organic food consumption of each food item. Organic 93 food intake was obtained for each item by applying a weight of 0, 0.25, 0.5, 0.75 and 1 to the five 94 respective categories of frequency (never, rarely, half of the time, often and always). Using the Org-95 FFQ, the proportion of organic food in the whole diet (g/d) was then calculated by dividing the total 96 organic food consumption out of the total food consumption excluding water. Participants were also 97 invited to fill in a questionnaire pertaining to motivations and attitudes towards organic foods.

98

99 Selection of the subsample and matching procedure

100 Of the 33,384 subjects who had completed the Org-FFQ, we selected those with available data 101 regarding total and organic food consumption, with no missing covariates and who had attended the 102 clinical visits (N=5,746). Among them, we selected subjects who had fasted at least 6 hours before the 103 visit and subjects who were no subject to potential metabolic disorders, i.e. subjects with no history of 104 type I diabetes, Crohn's disease, all types of cancer, neurological diseases, cardiovascular diseases, 105 digestive system diseases (including cirrhosis, hepatitis, celiac disease and colitis) lupus, 106 spondylolisthesis and sclerosis (N=4,598) (Figure 1). Then low and high organic food consumers 107 were identified as those with a proportion of organic food in the whole diet (g/d) below 10% or above 50%, respectively (N=2,351). Finally, in order to account for profile differences between low and high 108 109 organic food consumers, we applied a propensity score matching procedure without replacement. This 110 matching approach, based on a single composite score, enabled to achieve comparability between the two groups in terms of their observed characteristics ³³. We thus obtained two propensity score-111 matched groups of 150 subjects, differing by the organic valence of their diet. Selection and matching 112 113 procedures are extensively described in the Supplemental Material.

114

115 Urine collection, creatinine and pesticide analysis

At the clinical visit, urine sample collection was performed using vessels allowing the close-circuit 116 117 urine transfer from the vessel to the Vacutainer® tube. The Vacutainer® tubes containing the spot urine sample were kept at + 4°C before and during transportation to the central laboratory. After 118 119 splitting in aliquots, urine samples were stored at - 80°C for further analyses. To account for urine dilution, creatinine concentration (µg/L) was used to adjust analyte concentrations. Urinary creatinine 120 concentration was determined by ¹H NMR according to a method adapted from Bouatra et al. ³⁴. All 121 pesticide assays were performed in the same laboratory. The final list of analyzed pesticides and 122 123 metabolites has been defined as a compromise between scientific objectives, financial cost and the 124 available measurement methods. The extraction method for Group 1 (dimethylphosphate (DMP), dimethylthiophosphate (DMTP), dimethyldithiophosphate (DMDTP), diethylphosphate (DEP), 125 126 diethylthiophosphate (DETP), diethyldithiophosphate (DEDTP)) was 96-well µElution Solid Phase 127 Extraction-off line, Oasis Wax well Plate, 30 µm, 2 mg, Waters (Milford, Massachusetts, USA) for 128 purification and concentration purposes. The sample volume used was 200 µl, diluted by 200 µl of 129 water 4% formic acid. The compounds were eluted using a 5% ammonium hydroxide solution in 130 acetonitrile. The extraction for other analytes - Group 2 (chlorpyrifos and metabolites, malathion,

dichlorvos, phoxim, diazinon, thiabendazole-5-OH (TBZ-OH), tebuconazole, 2-(diethylamino)-6methylpyrimidin-4-ol/one (DEAMP) (pyrimiphos methyl metabolite), pyrethroid metabolites (3phenoxybenzoic acid (3-PBA), 4-fluoro-3-phenoxybenzoic acid (4-FPBA)) and 3,5,6trichloropyridinol (TCP)) was performed with Solid Phase Extraction-off line, Oasis HLB well Plate, 30 μ m, 2 mg, Waters (Milford, Massachusetts, USA). The sample volume used was 250 μ l, diluted by 250 μ l of water 4% ammonium hydroxide. The elution was done using a methanol-acetonitrile solution acidified with formic acid (0.05%).

A second analysis was performed for the deconjugated compounds before the same protocol of extraction: an enzymatic hydrolysis was performed, on 250 μ l sample, for total (free and conjugated) TBZ-OH, tebuconazole, DEAMP, pyrethroid metabolites (3-PBA et 4-FPBA) and TCP, using β -Glucuronidase / Arylsulfatase (Helix pomatia) from Roche (Mannheim, Germany), after stabilization with sodium acetate buffer solution, pH 4.5-5.5, as described by the supplier (Sigma-Aldrich, Seelze, Germany), during 16 hours. at 37°C.

Analyses of pesticides and metabolites (for correspondence between metabolites and parent compounds, see **Supplemental Table 1**) were performed using UPLC H-Class system coupled with a tandem mass spectrometry Xevo TQ-S (UPLC-MSMS) (Waters, Milford, USA).

For the pesticides from Group 1, a volume of 15 μ l was injected in a column BEH amide 1.7 μ m 2.1 x 100 mm (Waters, Milford, USA) in an oven at 35°C. The eluents used were water 50 mM ammonium acetate / acetonitrile with a gradient from 90% to 20% acetonitrile, with a flow rate of 0,4 ml/min. The analysis was performed in ES-mode. For the pesticides from Group 2, a volume of 10 μ l was injected in a column BEH C18 1.7 μ m 2.1 x 100 mm (Waters, Milford, USA). The eluents used were water/acetonitrile + 0.05 % formic acid with a gradient from. 85.5% to 2% water, in an oven at 40°C temperature with a flow rate of 0.4 ml/min The analysis was performed in ES + and ES- mode.

154

155 *Quality control/quality assurance*

156 All standards used were certified standards provided by Sigma-Aldrich (Seelze, Germany), A2S (St

157 Jean d'Illac, France) and Dr Ehrenstorfer (Augsburg, Germany).

A set of five calibrations in the range 0.01 to 10 μ g/L for samples of human urine free of the target 158 159 biomarkers was analyzed each day for 5 days for inter-assay precision and accuracy. For intra-assay precision and accuracy, three levels of concentrations (near limit of quantification, average and high 160 161 level) were analyzed in different human urine free of the target biomarkers ten times. A weighted linear regression model (1/x) was used for the construction of calibration curve. Throughout the study 162 a quality control was analyzed every ten samples and a blank every twenty samples. Dichlorvos-D6, 163 chlorpyrifos-methyl-D6, chlorpyrifos-D10, malathion-D6, diazinon-D10, DETP-D10 and DMTP-D6 164 165 served as internal standards.

The limit of detections (LODs) ranged from 0.003 to 0.6 μ g/L. The limit of quantification (LOQs) ranged from 0.01 to 2 μ g/L with a coefficient of variation ranging from 10 to 25%. LODs were 3 fold higher the intensity of the background noise and LOQs were, overall, 3 fold higher than LODs.

169

170 Statistical analysis

For comparative purposes and as previously done in other publications ^{18,21}, we calculated the following molar sums (μ mol/g), based on concentration of individual organophosphorus metabolites (μ g/g creatinine) and molecular weights (g/mol):

$$total MPs = \frac{[DMP]}{125} + \frac{[DMTP]}{141} + \frac{[DMDTP]}{157}$$
$$total EPs = \frac{[DEP]}{153} + \frac{[DETP]}{169} + \frac{[DEDTP]}{186}$$
$$total DAPs = total MPs + total EPs$$

with DMP, dimethylphosphate ; DMTP, dimethylthiophosphate ; DMDTP, dimethyldithiophosphate ;
DEP, diethylphosphate ; DETP, diethylthiophosphate ; DEDTP, diethyldithiophosphate ; DAPs,
dialkylphosphates ; MPs, dimethylphosphates ; EPs, diethylphosphates.

177 Samples containing concentrations below the LOD (limit of detection) were assumed to have 178 concentrations equal to $\frac{1}{2}$ LOD. Samples containing concentrations below the LOQ (limit of 179 quantification) were assumed to have a concentration equal to the midpoint between the LOD and the 180 LOQ ³⁵. 181 We performed a balance diagnostic of the matching procedure to obtain standardized differences for variables included in the propensity score model as recommended ³³, using the SAS macro %pmdiag 182 183 ³³. Characteristics of the participants are presented by group of consumers and were compared using 184 Wilcoxon signed-rank test for matched samples for continuous variables and McNemar test (binary variables) or conditional logistic regression for categorical variables (>2-class variables). We 185 computed the modified Programme National Nutrition Santé Guideline Score (mPNNS-GS), an a 186 priori nutritional index reflecting the adherence to the French food-based nutritional guidelines ³⁶. A 187 higher score (max=13.5) reflects a higher nutritional quality of the diet ³⁷. Distribution indicators, 188 189 frequency of detection and of quantification are provided.

Additionally, in a sensitivity analysis, in order to increase the discriminating power of our analyses, we only considered subjects with proportion of organic food in the diet below 5% for the conventional group and their matching organic pairs (n=218). All analyses were performed using 9.4 version of the SAS software (SAS Institute Inc., Cary, NC, USA).

194

195 Results

The balance diagnostic of the matching procedure is presented in **Supplemental Table 2**. The vast majority of the variables including health, sociodemographic and diet displayed similar distributions across organic and conventional groups, except for consumption of mixed dishes (p-value<0.05). Participants' characteristics are shown in **Table 1**. The average proportions of organic food in the diet were 3% (\pm 3) and 67% (\pm 13) in the conventional and organic groups, respectively. Diet quality, assessed using the mPNNS-GS, was relatively high. Mean age participants was 58y, about 30% of the participants were men and more than 60% were highly educated.

Levels of exposure to pesticides through urinary parent moieties and metabolites are presented in **Tables 2** and **3**. Parent pesticides were detectable in a limited number of samples in the two groups (max=9% for malathion in the conventional group). Mean and median molecule concentrations were mostly below the LOD (**Table 2**). In contrast, for some metabolites such as total DEAMP, total TBZ-OH, total tebuconazole, most EPs and MPs (except DEDTP and DMDTP) and free 3-PBA, the frequency of quantification was high in both groups (>15), and generally lower in the organic group (Table 3). The mean concentrations of DETP, DMTP and free 3-PBA were significantly higher in the
conventional group compared to the organic one while for the rest of metabolites, no significant
difference between groups was detected (Table 3).

The molar sums of EPs, MPs and DAPs are presented in **Figure 2** and **Table 4**. While no significant difference in urinary concentrations across groups was observed for total MPs (p-value=0.47) and total EPs (p-value=0.09), the total DAPs concentration was lower in the organic group compared to the conventional group, with means of 0.29 and 0.41 μ g/g creatinine respectively (p-value=0.03). Some consumers had high exposure levels in both groups as shown by rather different orders of magnitude of means and medians.

When the population sample was restricted to pairs using a different criteria, namely <5% of organic food in the diet for conventional consumers (instead of 10% as previously stated) (n=218), mean difference for total EPs was accentuated and reached statistical significance (**Table 5**), indicating a possible dose-response relationship.

222

223 Discussion

224 In this observational study, when comparing urine pesticide metabolites in consumers with discriminant consumption of organic food, significantly higher urinary levels of DETP, DMTP, total 225 226 DAPs (organophosphorus metabolites) and free 3-PBA (a pyrethroid metabolite) were found among 227 conventional consumers compared to organic consumers, while, overall, low detection rates were 228 found for parent compounds in both groups. In a sensitivity analysis, with conventional consumers 229 with a maximum of 5% of organic food in their diet, differences between groups were more salient, in 230 particular for EPs. This suggests that organophosphorus and pyrethroid pesticide exposure in adults 231 may be noticeably lowered by introducing organic foods in the diet.

During the last decade, several studies comparing the levels of urine pesticides between "organic" and "conventional" consumers have been conducted in children ^{17–19,38} and in adults ^{20,21}. Beyond disparities in age range and dietary patterns of the populations as well as the periods when they have been conducted, these studies largely differed in their methodology. Most of them were indeed interventional cross-over studies ^{17,19,21,38}, two were observational ^{18,20}. An exposure to a wide range of 237 metabolites was assessed including organophosphorus pesticide metabolites but also, less frequently, 238 pyrethroids and some herbicides. Studies conducted among children reported reduction in total DAPs ^{17,18} or metabolites of malathion and chlorpyrifos ^{19,38} after switching to an organic controlled diet or 239 240 when comparing organic and conventional participants. A drastic and immediate reduction was reported in cross-over studies. For instance, among 40 Mexican-American 3-6 years children, total 241 DAPs reduction was 40% after 7 days of organic diet ¹⁷. Similarly, in a study conducted among 242 preschool children, mean DMP urinary concentration was 9 times lower among 18 children following 243 an organic diet compared to 21 children eating a conventional one ¹⁸. Overall, these studies argue for a 244 245 central role of the mode of production of food in organophosphorus pesticide exposure among 246 children. These results were observed even in the study with an observational design reflecting actual levels of organic food consumption ¹⁸. 247

In the present study conducted in French adults, significant reductions in median urinary concentrations - ranging from 17% to 55% - were observed for DETP, DMTP and free 3-PBA in organic consumers compared to conventional consumers while no significant results were found for the other pesticides investigated.

252 To the best of our knowledge, only one experimental crossover study has been carried out among adults to test for modification in urinary pesticide metabolites following a period of organic diet ²¹. In 253 that crossover study conducted among 13 adults in Australia, participants alternated two 7-day periods 254 255 during which diets were either composed of at least 80% of conventional foods or organic foods. All 256 DAPs were less frequently detected during the organic period than during the conventional one. After 257 applying the same calculation as in our study for left-censored data, a reduction of 89% of total DAPs was observed. Indeed, during the organic phase, total DAPs mean concentration was 0.032 ± 0.038 258 μ mol/g creatinine while total DAPs mean concentration among "organic consumers" was 0.29 ± 0.42 259 260 in our study. During the conventional diet period, total DAPs mean concentration was 0.294 ± 0.435 μ mol/g creatinine while we found total DAPs mean concentration of 0.41 ± 0.64 among conventional 261 262 consumers. These differences in the level of exposure may be explained by different designs 263 (experimental vs observational) as well as by different methodological approaches. Indeed, our study 264 is based on observational data, i.e. based on actual self-reported data with participants that did not

265 consume 100% organic or conventional food. In other words, given the cut-off used herein (i.e. 266 minimum 50% of the diet coming from organic sources) conventional foods could also make up a 267 quite important part of the diet in the "organic group", leading to a certain and variable dietary 268 pesticide exposure. In the same line, our sensitivity analysis data clearly showed that reducing the proportion of organic foods (from maximum 10% to 5%) in the conventional consumer diet 269 270 exacerbated the differences between groups in some metabolite concentrations. In addition, in our 271 study, the definition of organic consumers was based on data collected using a self-administered food 272 frequency questionnaire covering the past year. Use of a self-administered food frequency 273 questionnaire, prone to some measurement error may have led to misclassification. Moreover, as organophosphorus metabolites exhibit a short half-live ³⁹, it is possible that some participants 274 identified as organic consumers have consumed highly contaminated (conventional) food just before 275 the urine sampling. As underlined in another study 20 , the very short half live of these compounds may 276 strongly limit the relevance of this type of measure to assess the overall exposure to 277 278 organophosphorus.

In the only other observational study carried out in adults, urinary DAPs and self-reported organic food consumption habits (using a 3-categories question) were assessed in 480 US participants ²⁰. In that study, total DAPs concentrations significantly decreased while increasing consumption of organic food with a difference of 65% between never organic food consumers and often/always organic food consumers. The magnitudes of the effects of switching from conventional to organic diets were comparable to those observed in our study, i.e. lower than in experimental controlled trials.

Herein, total DEAMP, total TBZ-OH, total tebuconazole, most EPs and MPs, and free 3-PBA were frequently detected with rather high levels, even among organic consumers. As already mentioned, this may be explained by the fact that, in so-called high organic food consumers, organic food made up, on average, "only" 63% of the diet, meaning that, on average, almost 40% of the food was conventional. These individuals may have also been big consumers of conventional fruit and vegetables, leading to an overall quite high dietary pesticide exposure.

291 Compared with pesticide exposures estimated in 2006 in a random representative French survey 292 (ENNS), participants of the conventional group in our study exhibited comparable urinary pesticide

levels for DETP, DMTP and lower for DMDTP ⁴⁰. In contrast, mean urinary concentrations of DEP 293 and DMP were markedly higher in our study ⁴⁰. For instance, mean and median of DEP urinary 294 295 concentrations among our conventional group were 31.68 and 0.96 μ g/g creatinine while in the ENNS 296 study the respective values were 3.89 and 3.66. These findings reveal extremely different distributions of pesticide exposure levels between the ENNS survey and our study in which some participants had 297 298 very high exposure levels. These differences can also partly be explained by the different approach 299 used for left-censored data. Concerning pyrethroid metabolites, the median urinary concentration of 300 free 4-F-3-PBA was very low in our study (LOD= $0.02 \mu g/L$), as in the ENNS study (<LOD=0.03301 μ g/L). For 3-PBA, the urinary concentration was higher in the ENNS study with a geometric mean of 302 $0.72 \,\mu g/g$ creatinine which compares to a mean of 0.13 $\mu g/g$ creatinine in the conventional group of 303 our study. As we mentioned above, it can be hypothesized that since subjects in our sample were very high consumers of fruit, vegetables and whole grains compared to the French population ⁴¹, they must 304 have been particularly exposed to pesticides from plant-based products, which are indeed largely 305 sprayed by synthetic chemicals ¹⁶. This hypothesis is reinforced by the observation that some urinary 306 307 organophosphorus pesticide metabolites concentrations (namely DETP and DMTP) and 3-PBA are 308 linearly associated with conventional fruit and vegetable consumption (data not shown) in our study. For instance a factor 4 was observed for DETP urinary concentration between 1st and 3rd tertiles of 309 310 consumption of conventional fruit and vegetables.

311 Some limitations should be highlighted. First, our study design is entirely observational and based on self-reported data. Thus, consumption data are prone to misestimating, and in particular organic food 312 consumption may have been overestimated ³¹. While food consumption data are relatively precisely 313 314 recorded, we did not quantify the extent of other sources of exposure (i.e. occupational, home or 315 environmental ones). For instance, thiabendazole is not only authorized as a plant protection product 316 but is also a biocide for wood treatment (see **Supplemental Table 1**). In addition, it is a preservative 317 authorized for use in foodstuffs intended for human consumption (E233), and an anthelminthic drug 318 used in human and veterinary medicine. In this case, the presence of thiabendazole metabolites in 319 urine can be due to multiple non-dietary sources, although it should be born in mind that diet remains 320 the main source of pesticide exposure in the general population. The absence of difference in 321 pesticides exposure may also be due to a lack of power, at least for contaminants with very low 322 detection or quantification levels. The use of a middle bound scenario for left-censored data may also 323 have led to the overestimation of some estimates.

324 Another limit pertains to the efficiency of matching. Although a wide range of confounders were used 325 to estimate the propensity score, we cannot rule out possible residual confounding between organic 326 and conventional consumers. In addition, propensity score matching leads to the exclusion of "particular" subjects unable to be matched, thus avoiding the external validity of findings as 327 previously highlighted ³³. Our study sample, composed of volunteers, was particular with respect to 328 329 sociodemographic characteristics and dietary intakes and thus is not comparable to the general population. NutriNet-Santé participants have indeed higher intakes of fruit and vegetables than the 330 general French population ⁴¹. Finally, as previously mentioned, concerning urinary DAPs and 3-PBA 331 concentrations, the reliability of such biomarkers to reflect long-term exposure (usual diet, as assessed 332 with the Org-FFQ) is questionable as their half live is short ^{39,42} and only one biological sampling has 333 been available. It should also be noted that given most pesticide moieties studied herein were no 334 335 longer authorized for use in conventional agriculture at the time of urine sampling, the differences in 336 overall pesticide exposure between the two groups of consumers are likely under-estimated.

337

Our study also exhibits major strengths. Detailed data was used to assess organic and conventional consumption in the overall diet. We used an effective method, i.e. propensity score, allowing to match organic and conventional consumers using a wide range of covariates (including sociodemographic, dietary and health data). Finally, this is the first study conducted in Europe comparing pesticide urinary concentrations from different classes of pesticides, in adults who differed by their organic consumption in real conditions.

344

345 Conclusions

Compared to individuals with low organic food consumption, individuals with high proportion of organic food in their diet had significantly lower levels of various metabolites of pesticides of the organophosphate and pyrethroid families, suggesting that an organic food based-diet may help reduce 349 the dietary pesticide exposure, at least for some agrochemicals as tested herein. Overall, low detection rates were found for parent compounds in both groups. It also should be stressed that urines of 350 351 participants in our study displayed rather high exposure levels, irrespective of the group considered, 352 compared to other populations.. It would be also of high interest to conduct similar comparisons in 353 large and representative samples. Accurate assessment of organic food consumption in representative 354 national surveys is therefore required to better evaluate the clinical effects of such differences at the national level. Additional research is also needed to assess the part attributable to dietary pesticide 355 356 exposure and non-dietary sources as well as the effects of such differences over the long term on 357 health status.

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- 361

362 List of abbreviations:

- 363 3-PBA, 3-phenoxybenzoic acid;
- 364 4-F-3-PBA, 4-fluoro-3-phenoxybenzoic acid;
- 365 CPMO, chlorpyrifos-methyl-oxon;
- 366 CPO, chlorpyrifos-oxon
- 367 DAPs, dialkylphosphates;
- 368 DEAMP, 2-(diethylamino)-6-methylpyrimidin-4-ol/one;
- 369 DEDTP, diethyldithiophosphate;
- 370 DEP, diethylphosphate;
- 371 DETP, diethylthiophosphate;
- 372 DMDTP, dimethyldithiophosphate;
- 373 DMP, dimethylphosphate;
- 374 DMTP, dimethylthiophosphate;
- 375 ENNS, Etude Nationale Nutrition Santé (French Nutrition and Health Survey)
- 376 EPs, diethylphosphates;
- 377 F-PBA, 4-fluoro-3-phenoxybenzoic acid;
- 378 GC–MS/MS, gas chromatography tandem mass spectrometry;
- 379 LOD, limit of detection;
- 380 LOQ, limit of quantification;
- 381 mPNNS-GS, modified Progamme National Nutrition Santé Guideline Score;
- 382 MPs, dimethylphosphates;
- 383 Org-FFQ, organic food frequency questionnaire;
- 384 PBA, 3-phenoxybenzoic acid;
- 385 TBZ-OH, 5-hydroxy-thiabendazole;

387

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391

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400

401 Authors' contributions

The author contributions were as follows: JPC, SH, EKG, DL and PG conceived and designed the experiments. JB performed the experiments. GD and AD performed the urinary dosages. JB and EKG analyzed the data. JB and EKG wrote the paper. All the authors were involved in interpreting results and editing the manuscript. JB and EKG had primary responsibility for final content. All authors read and approved the final manuscript.

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- 519

| | Conventional group | Organic group | P-value ^b |
|---|--------------------|-------------------|----------------------|
| Ν | 150 | 150 | |
| Proportion (%) of organic food in the diet | 0.03 ± 0.03 | 0.67 ± 0.13 | <0.0001* |
| Age (y) | 58.71 ± 12.78 | 58.35 ± 11.69 | 0.60 |
| % Male | 28% | 32% | 0.47 |
| Energy intake (kcal/d) | 1927 ± 561.3 | 1994 ± 601.6 | 0.37 |
| mPNNS-GS ^c (13.5) | 8.73 ± 1.76 | 8.73 ± 1.67 | 0.59 |
| Body mass index (BMI) (kg/m²) | 24.18 ± 4.11 | 24.19 ± 4.02 | 0.93 |
| Tobacco status (%) | | | 0.38 |
| never smoker | 48.67 | 47.33 | |
| former smoker | 44 | 40.67 | |
| current smoker | 7.33 | 12 | |
| Physical activity (%) | | | 0.85 |
| missing | 9 | 8 | |
| low | 17 | 19 | |
| medium | 37 | 35 | |
| high | 37 | 38 | |
| Vegetarian or vegan diet (yes) | 1.33 | 2 | 0.65 |
| (%) | 1.55 | 2 | |
| Location (%) | | | 0.75 |
| rural community | 24 | 25.33 | |
| urban unit with a population | 11.33 | 14 | |
| smaller than 20 000 inhabitants | 11.55 | 17 | |
| urban unit with a population | | | |
| between 20 000 and 200 000 | 18.67 | 20.67 | |
| inhabitants | | | |
| urban unit with a population | 46 | 40 | |
| higher than 200 000 inhabitants | 40 | U | |
| Education (%) | | | 0.60 |
| < high school diploma | 22 | 22 | |
| high school diploma | 12.67 | 16.67 | |
| > high school diploma | 65.34 | 61.34 | |
| Monthly income per household | | | 0.77 |
| unit (€)% | | | 0.77 |
| refused to declare | 9.33 | 7.33 | |
| 900-1200 | 6.67 | 8.67 | |
| 1200-1800 | 22.67 | 20 | |
| 1800-2700 | 22 | 26.67 | |
| >2700 | 39.33 | 37.33 | |

Table 1: Main characteristics of the sample, n=300, NutriNet-Santé^a

Abbreviation: mPNNS-GS, modified Programme National Nutrition Santé Guidelines score

^aMeans (± SD) or percentages as appropriate.

^bP-values referred to Wilcoxon matched pair signed-rank tests for continuous variables, Mc Nemar's

test or conditional logistic regression for categorical variables

^cmPNNS-GS is a dietary index reflecting the level of adherence to French nutritional guidelines

| Pesticides (µg/g creatinine) | Mean \pm SD | P10 | P50 | P90 | P- value ^a | LOD (µg/L) | LOQ (µg/L) | Crude mean (µg/L) | Crude median (µg/L) | % of detection | % of quantificatio n |
|---------------------------------|---------------------|--------|--------|--------|--------------------------|---------------|-------------------|-------------------------|------------------------|----------------|----------------------------|
| Malathion | | | | | 0.09 | 0.003 | 0.01 | | | | |
| Organic group | 0.0083 ± 0.0321 | 0.0009 | 0.0020 | 0.0068 | | | | 0.0039 | 0.0015 | 3 | 3 |
| Conventional group | 0.0080 ± 0.0283 | 0.0009 | 0.0023 | 0.0116 | | | | 0.0041 | 0.0015 | 9 | 7 |
| Chlorpyrifos | | | | | 0.55 | 0.02 | 0.05 | | | | |
| Organic group | 0.0405 ± 0.2419 | 0.0058 | 0.0138 | 0.0420 | | | | 0.0151 | 0.0100 | 3 | 2 |
| Conventional group | 0.0227 ± 0.0287 | 0.0056 | 0.0144 | 0.0511 | | | | 0.0103 | 0.0100 | 2 | 0 |
| Chlorpyrifos- methyl | | | | | 0.45 | 0.2 | 0.5 | | | | |
| Organic group | 0.1886 ± 0.1520 | 0.0582 | 0.1350 | 0.3975 | | | | 0.1010 | 0.1000 | 1 | 1 |
| Conventional group | 0.2089 ± 0.1899 | 0.0555 | 0.1437 | 0.4858 | | | | 0.100 | 0.1000 | 0 | 0 |
| Phoxim | | | | | 0.65 | 0.05 | 0.1 | | | | |
| Organic group | 0.0510 ± 0.0439 | 0.0148 | 0.0348 | 0.1089 | | | | 0.0298 | 0.0250 | 5 | 2 |
| Conventional group | 0.0569 ± 0.0643 | 0.0139 | 0.0371 | 0.1246 | | | | 0.0272 | 0.0250 | 3 | 2 |
| Diazinon | | | | | 0.33 | 0.02 | 0.05 | | | | |
| Organic group | 0.0188 ± 0.0152 | 0.0058 | 0.0133 | 0.0398 | | | | 0.0100 | 0.0100 | 0 | 0 |
| Conventional group | 0.0277 ± 0.0654 | 0.0056 | 0.0144 | 0.0511 | | | | 0.0124 | 0.0100 | 1 | 1 |
| Dichlorvos | | | | | 0.42 | 0.3 | 0.9 | | | | |
| Organic group | 0.2819 ± 0.2285 | 0.0873 | 0.1993 | 0.5962 | | | | 0.1500 | 0.1500 | 0 | 0 |
| Conventional group | 0.3134 ± 0.2849 | 0.0833 | 0.2156 | 0.7286 | | | | 0.1500 | 0.1500 | 0 | 0 |

Table 2: Results for the pesticide parent compounds across organic (n=150) and conventional (n=150) groups, NutriNet-Santé

Abbreviations: LOD, Limit of detection; LOQ, Limit of quantification

^aP-values referred to Wilcoxon matched pair signed-rank tests

| Pesticides (µg/g creatinine) | Mean \pm SD | P10 | P50 | P90 | P-value ^a | LOD (µg/L) | LOQ (µg/L) | Crude mean (µg/L) | Crude median (µg/L) | % of detection | % of quantification |
|---------------------------------|----------------------|--------|--------|--------|----------------------|---------------|---------------|----------------------|------------------------|----------------|---------------------|
| СРО | | | | | 0.45 | 0.005 | 0.01 | | | | |
| Organic group | 0.0050 ± 0.0053 | 0.0015 | 0.0033 | 0.0102 | | | | 0.0026 | 0.0025 | 1 | 1 |
| Conventional group | 0.00548 ± 0.0062 | 0.0014 | 0.0036 | 0.0125 | | | | 0.0025 | 0.0025 | 1 | 0 |
| СРМО | | | | | 0.42 | 0.02 | 0.05 | | | | |
| Organic group | 0.0188 ± 0.0152 | 0.0058 | 0.0133 | 0.0398 | | | | 0.0100 | 0.0100 | 0 | 0 |
| Conventional group | 0.0209 ± 0.0190 | 0.0056 | 0.0144 | 0.0486 | | | | 0.0100 | 0.0100 | 0 | 0 |
| free TCP | | | | | 0.32 | 0.2 | 0.5 | | | | |
| Organic group | 0.1918 ± 0.1574 | 0.0582 | 0.1350 | 0.4097 | | | | 0.1020 | 0.1000 | 1 | 0 |
| Conventional group | 0.2254 ± 0.2337 | 0.0610 | 0.1485 | 0.5110 | | | | 0.1183 | 0.1000 | 3 | 1 |
| total TCP | | | | | 0.14 | 0.2 | 0.5 | | | | |
| Organic group | 0.3177 ± 0.4123 | 0.0590 | 0.1794 | 0.7078 | | | | 0.1804 | 0.1000 | 21 | 8 |
| Conventional group | 0.4155 ± 0.6757 | 0.0678 | 0.2029 | 0.7902 | | | | 0.2435 | 0.1000 | 29 | 13 |
| free DEAMP | | | | | 0.46 | 0.03 | 0.1 | | | | |
| Organic group | 0.0404 ± 0.0620 | 0.0090 | 0.0237 | 0.0733 | | | | 0.0232 | 0.0150 | 10 | 5 |
| Conventional group | 0.0378 ± 0.0379 | 0.0091 | 0.0230 | 0.0798 | | | | 0.0212 | 0.0150 | 8 | 3 |
| total DEAMP | | | | | 0.06 | 0.03 | 0.1 | | | | |
| Organic group | 0.6121 ± 1.368 | 0.0115 | 0.1104 | 1.386 | | | | 0.3927 | 0.0500 | 65 | 42 |
| Conventional group | 0.6953 ± 1.243 | 0.0151 | 0.2116 | 1.894 | | | | 0.4874 | 0.1400 | 70 | 58 |

Table 3: Results for the pesticide metabolites across organic (n=150) and conventional (n=150) groups, NutriNet-Santé

| free TBZ-OH | | | | | 0.60 | 0.03 | 0.1 | | | | |
|--------------------|---------------------|--------|--------|--------|---------|------|-----|--------|--------|----|----|
| Organic group | 0.0339 ± 0.0445 | 0.0087 | 0.0209 | 0.0654 | | | | 0.0181 | 0.0150 | 5 | 1 |
| Conventional group | 0.0336 ± 0.0315 | 0.0087 | 0.0225 | 0.0767 | | | | 0.0180 | 0.0150 | 3 | 1 |
| total TBZ-OH | | | | | 0.26 | 0.03 | 0.1 | | | | |
| Organic group | 0.2762 ± 0.8386 | 0.0095 | 0.0424 | 0.5728 | | | | 0.1664 | 0.0150 | 38 | 26 |
| Conventional group | 0.1635 ± 0.2193 | 0.0124 | 0.0757 | 0.4165 | | | | 0.1324 | 0.0500 | 51 | 35 |
| free tebuconazole | | | | | 0.44 | 0.03 | 0.1 | | | | |
| Organic group | 0.0292 ± 0.0240 | 0.0087 | 0.0206 | 0.0630 | | | | 0.0157 | 0.0150 | 2 | 0 |
| Conventional group | 0.0334 ± 0.0364 | 0.0083 | 0.0217 | 0.0748 | | | | 0.0158 | 0.0150 | 1 | 1 |
| total tebuconazole | | | | | 0.83 | 0.03 | 0.1 | | | | |
| Organic group | 0.2859 ± 0.5257 | 0.0136 | 0.0937 | 0.7685 | | | | 0.2554 | 0.0500 | 68 | 37 |
| Conventional group | 0.2704 ± 0.5995 | 0.0153 | 0.0904 | 0.6416 | | | | 0.1986 | 0.0500 | 61 | 41 |
| DEP | | | | | 0.14 | 0.2 | 0.6 | | | | |
| Organic group | 18.61 ± 42.06 | 0.0646 | 0.6388 | 55.01 | | | | 10.61 | 0.1000 | 49 | 48 |
| Conventional group | 31.68 ± 69.38 | 0.0684 | 0.9682 | 80.96 | | | | 15.91 | 0.6800 | 52 | 51 |
| DETP | | | | | 0.0003* | 0.2 | 0.6 | | | | |
| Organic group | 0.4305 ± 1.208 | 0.0623 | 0.1963 | 0.8601 | | | | 0.2775 | 0.1000 | 22 | 11 |
| Conventional group | 1.018 ± 2.480 | 0.0758 | 0.2966 | 1.966 | | | | 0.6775 | 0.1000 | 37 | 24 |
| DEDTP | | | | | 0.42 | 0.2 | 0.6 | | | | |
| Organic group | 0.1879 ± 0.1523 | 0.0582 | 0.1329 | 0.3975 | | | | 0.1000 | 0.1000 | 0 | 0 |

| Conventional group | 0.2089 ± 0.1899 | 0.0555 | 0.1437 | 0.4858 | | | | 0.1000 | 0.1000 | 0 | 0 |
|--------------------|---------------------|--------|--------|--------|--------|------|------|--------|--------|----|----|
| DMP | | | | | 0.49 | 0.6 | 2 | | | | |
| Organic group | 17.85 ± 35.22 | 0.2253 | 5.427 | 50.61 | | | | 14.30 | 3.860 | 65 | 61 |
| Conventional group | 18.66 ± 45.14 | 0.2171 | 3.132 | 40.62 | | | | 11.89 | 2.330 | 61 | 53 |
| DMTP | | | | | 0.001* | 0.2 | 0.6 | | | | |
| Organic group | 2.654 ± 6.488 | 0.0894 | 0.6209 | 6.603 | | | | 2.103 | 0.3000 | 62 | 44 |
| Conventional group | 6.310 ± 18.55 | 0.1676 | 1.382 | 13.88 | | | | 4.102 | 1.245 | 73 | 62 |
| DMDTP | | | | | 0.25 | 0.2 | 0.6 | | | | |
| Organic group | 0.1879 ± 0.1523 | 0.0582 | 0.1329 | 0.3975 | | | | 0.1000 | 0.1000 | 0 | 0 |
| Conventional group | 0.2212 ± 0.2080 | 0.0555 | 0.1437 | 0.5189 | | | | 0.1053 | 0.1000 | 3 | 0 |
| | | | | | | | | | | | |
| free 3-PBA | | | | | 0.03* | 0.02 | 0.05 | | | | |
| Organic group | 0.0380 ± 0.0730 | 0.0061 | 0.0171 | 0.0610 | | | | 0.0263 | 0.0100 | 16 | 10 |
| Conventional group | 0.0579 ± 0.1023 | 0.0068 | 0.0208 | 0.1137 | | | | 0.0415 | 0.0100 | 26 | 21 |
| total 3-PBA | | | | | 0.16 | 0.02 | 0.05 | | | | |
| Organic group | 0.0885 ± 0.2877 | 0.0064 | 0.0201 | 0.1798 | | | | 0.1178 | 0.0100 | 23 | 18 |
| Conventional group | 0.1301 ± 0.3536 | 0.0076 | 0.0282 | 0.3429 | | | | 0.1303 | 0.0100 | 35 | 29 |
| | | | | | | | | | | | |
| free 4-F-3-PBA | | | | | 0.32 | 0.02 | 0.05 | | | | |
| Organic group | 0.0191 ± 0.0154 | 0.0059 | 0.0133 | 0.0410 | | | | 0.0102 | 0.0100 | 1 | 0 |
| Conventional group | 0.0222 ± 0.0221 | 0.0060 | 0.0149 | 0.0499 | | | | 0.0122 | 0.0100 | 3 | 1 |
| total 4-F-3-PBA | | | | | 0.71 | 0.02 | 0.05 | | | | |
| Organic group | 0.0237 ± 0.0392 | 0.0059 | 0.0137 | 0.0447 | | | | 0.0123 | 0.0100 | 3 | 2 |
| | | | | | | | | | | | |

Conventional group 0.0223 ± 0.0226 0.0060 0.0149 0.0499 0.0144 0.0100 3 1 Abbreviations: 3-PBA, 3-phenoxybenzoic acid; 4-F-3-PBA, 4-fluoro-3-phenoxybenzoic acid; CPMO, Chlorpyrifos-methyl-oxon; CPO, Chlorpyrifos-oxon; DAPs, dialkylphosphates; DEAMP, 2-(diethylamino)-6-methylpyrimidin-4-ol/one; DEDTP, diethyldithiophosphate; DEP, diethylphosphate; DETP, diethylthiophosphate; DMDTP, dimethyldithiophosphate; DMP, dimethylphosphate; DMTP, dimethylthiophosphate; EPs, diethylphosphates; LOD, Limit of detection; LOQ, Limit of quantification; MPs, dimethylphosphates; TBZ-OH, thiabendazole-5-OH; TCP, 3,5,6- trichloropyridinol ^aP-values referred to Wilcoxon matched pair signed-rank tests

| | Mean \pm SD | P10 | P50 | P90 | P-value ^b |
|--------------------|-----------------|------|------|------|----------------------|
| total EPs | | | | | |
| Organic group | 0.13 ± 0.28 | 0.00 | 0.01 | 0.36 | 0.09 |
| Conventional group | 0.21 ± 0.45 | 0.00 | 0.02 | 0.54 | |
| total MPs | | | | | 0.47 |
| Organic group | 0.16 ± 0.29 | 0.00 | 0.06 | 0.45 | |
| Conventional group | 0.20 ± 0.39 | 0.01 | 0.06 | 0.42 | |
| total DAPs | | | | | 0.03* |
| Organic group | 0.29 ± 0.42 | 0.01 | 0.12 | 0.82 | |
| Conventional group | 0.41 ± 0.64 | 0.01 | 0.16 | 1.23 | |
| | | | | | |

Table 4: Molar sum of DAP metabolites in μ mol/g creatinine organic (n=150) and conventional

(n=150) groups, NutriNet-Santé^a

Abbreviations: DAPs, dialkylphosphates; EPs, diethylphosphates; MPs, dimethylphosphates

^aConventional consumers were defined as individuals with a maximum of 10% of organic food in their

diet

^bP-values referred to Wilcoxon matched pair signed-rank tests

| | Mean \pm SD | P10 | P50 | P90 | P^{b} |
|--------------------|-----------------|------|------|------|---------|
| total EPs | | | | | 0.02* |
| Organic group | 0.11 ± 0.24 | 0.00 | 0.01 | 0.37 | |
| Conventional group | 0.25 ± 0.51 | 0.00 | 0.03 | 0.88 | |
| total MPs | | | | | 0.50 |
| Organic group | 0.17 ± 0.32 | 0.00 | 0.06 | 0.43 | |
| Conventional group | 0.20 ± 0.42 | 0.01 | 0.05 | 0.51 | |
| total DAPs | | | | | 0.04* |
| Organic group | 0.28 ± 0.42 | 0.01 | 0.11 | 0.84 | |
| Conventional group | 0.45 ± 0.70 | 0.01 | 0.17 | 1.37 | |

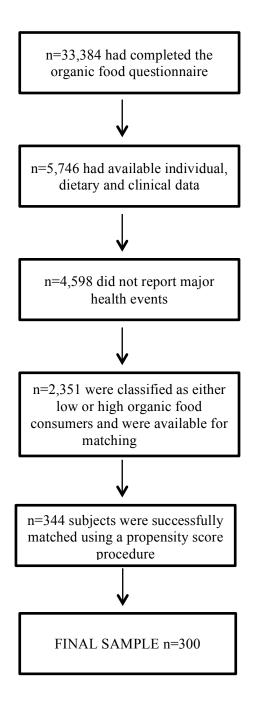
Table 5: Molar sum of DAP metabolites in μ mol/g creatinine across organic (n=109) and conventional (n=109) groups, NutriNet-Santé^a

Abbreviations: DAPs, dialkylphosphates; EPs, diethylphosphates; MPs, dimethylphosphates

^aCompared to data presented in Table 4, the population sample was restricted to pairs with

conventional consumers defined as individuals with a maximum of 5% of organic food in their diet

^bP-values referred to Wilcoxon matched pair signed-rank tests



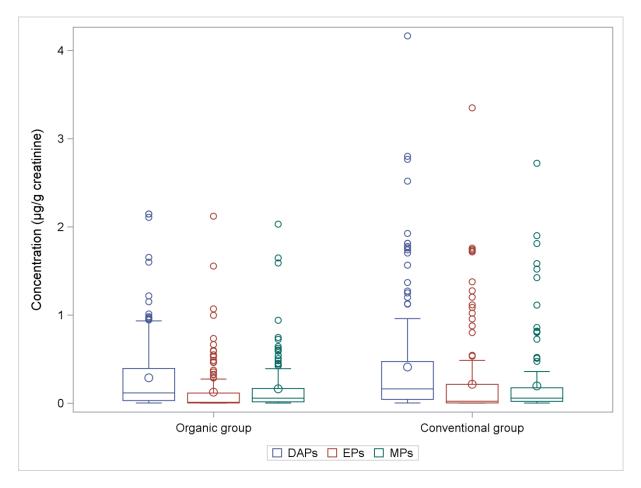


Figure 2: Molar sum of DAP metabolites across organic and conventional groups, n=300, NutriNet-

Santé

Abbreviations: DAPs, dialkylphosphates; EPs, diethylphosphates; MPs, dimethylphosphates