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# Important effects of relative humidity on the formation processes of iodine oxide particles from CH<sub>3</sub>I photo-oxidation

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## Abstract

In this work, laboratory chamber experiments of gas-phase methyl iodide photolysis in the presence of ozone at three relative humidity conditions were performed to study the formation and physico-chemical properties of iodine oxide particles. The obtained results revealed significant morphological changes of iodine oxide particles that were observed to depend on relative humidity. The formed iodine oxide particles under dry condition were supposed to be agglomerates of fine hygroscopic crystals. On the other hand, a humid atmosphere was observed to favor the formation of isomeric, tetragonal and orthorhombic hygroscopic crystals potentially composed of HIO<sub>3</sub> likely formed from progressive hydration of iodine oxide clusters. This process leads to a release of molecular iodine, I<sub>2</sub>, which may indicate a potential role of I<sub>2</sub>O<sub>4</sub> in the particles' evolution processes. The obtained results on the iodine oxides' behavior are important to the nuclear power plant safety industry since many of the organic iodides that may be released during a major nuclear power-plant accident contain radioactive isotopes of iodine that are known to have lethal or toxic impacts on human health.

**Keywords:** Iodine oxides, Iodides, Aerosols, Particles, Particle formation, Methyl iodide, Mass Spectrometry, SMPS, Granulometry, Microscopy, Electron microscopy

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## 1. Introduction

Organic iodides are of great interest for both the atmospheric science community and the nuclear power plant safety industry. They are involved in the chemical processes driving the oxidizing capacity of the atmosphere [1] and atmospheric new particle formation involved in marine aerosol and cloud formation [2]. This is important since atmospheric aerosols and clouds are known to scatter incoming solar radiation, thus, regulating the Earth's radiation budget [2,

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7 3]. Further, organic iodides are known to play a significant role in the tropospheric [4, 5] and  
8 stratospheric [6] ozone depletion cycles.

9 Organic iodides have also gained a considerable interest in the nuclear power plant safety  
10 industry following three major power plant accidents that include Three Mile Island (U.S.A.)  
11 [7], Chernobyl (Ukraine) [8] and Fukushima (Japan) [9] when the nuclear safety engineering  
12 community has come to understand that many different organic iodides are formed within the  
13 nuclear power plant containment building during such an accident [10, 11]. These newly formed  
14 organic iodides may then be released into the environment if the reactor building containment is  
15 breached. Further, following the Fukushima nuclear power plant accident, atmospheric transport  
16 models have clearly shown the importance to understand the atmospheric emissions and behav-  
17 ior of iodine species released from point sources [12, 13]. This is important, since many of these  
18 newly released organic iodides contain radioactive isotopes of iodine, namely  $^{131}\text{I}$  and  $^{129}\text{I}$ , that  
19 are known to have a lethal or toxic impact of human health. Primary compounds released dur-  
20 ing a major nuclear power plant accident include molecular iodine and methyl iodide ( $\text{CH}_3\text{I}$ ).  
21 The  $\text{CH}_3\text{I}$  is thought to be formed from the interactions of gas-phase iodine with the organic  
22 content present within the containment paint ingredients under gamma irradiation [14, 15, 16].  
23 The formed  $\text{CH}_3\text{I}$  is subsequently photolyzed by gamma or solar radiation in the UV region of  
24 the solar spectrum to produce  $\cdot\text{CH}_3$  and  $\cdot\text{I}$  radicals. The latter radicals are known to react with  
25 ozone (reaction R1), an oxidizing molecule readily available in ample quantities in the contain-  
26 ment area of a nuclear power plant following a major accident, to produce  $\cdot\text{IO}$  radicals, a known  
27 precursor of iodine oxide aerosols [17, 18].



28 The gas-phase  $\cdot\text{IO}$  radical's self-reaction has four potential thermodynamically accessible reac-  
29 tion channels that include the formation of  $\text{OIO} \cdot$  and  $\text{I} \cdot$  radicals (reaction R2a), the formation  
30 of  $\text{I}_2\text{O}_2$  (reaction R2b), the formation of  $\text{I} \cdot$  radical and molecular oxygen (reaction R2c) and the  
31 formation of molecular iodine and molecular oxygen (reaction R2d) [19, 20, 21, 22]:

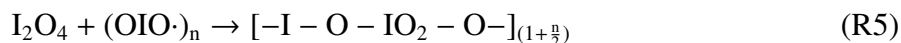




Only the reactive channel R2a that leads to the formation of  $\text{OIO} \cdot$  and  $\text{I} \cdot$  radicals is considered to be important and the reactive channels R2b to R2d are considered to be negligible or recycled further to regenerate  $\text{IO} \cdot$  radicals [23, 20, 24]. It is understood that the gas-phase  $\text{OIO} \cdot$  and  $\text{IO} \cdot$  radicals react further to form  $\text{I}_2\text{O}_4$  or  $\text{I}_2\text{O}_3$  (reactions R3 and R4) [23].



While studying the gas-phase  $\text{CH}_2\text{I}_2/h\nu/\text{O}_3$  reactivity in an atmospheric simulation chamber, Hoffmann et al. (2001) [17] suggested that the self-reaction of  $\cdot\text{OIO}$  (reaction R3) that leads to  $\text{I}_2\text{O}_4$  formation is the key nucleation precursor step responsible for aerosol formation. These authors' deduction is based on direct identification of  $\text{I}_2\text{O}_4$  molecules present within the formed aerosols and propose the following polymerization mechanism (reactions R3 and R5) to explain the nucleation process:



This observation is supported by theoretical and chamber work of Jimenez et al. (2003) who studied iodine oxide particle formation using the same reactive system as Hoffmann et al. (2001), namely  $\text{CH}_2\text{I}_2/h\nu/\text{O}_3$ , carried out under two different relative humidity experimental conditions, namely, dry and  $\approx 65\% \text{RH}$  [18]. Jimenez et al. (2003) [18] argued that the self-reaction of  $\cdot\text{OIO}$  radicals leads to the formation of  $\text{I}_2\text{O}_4$ , a conclusion that was based on the observed hygroscopic behavior of aerosols. These investigators argued that low hygroscopicity aerosols are best explained by the nucleation process involving  $\text{I}_2\text{O}_4$  and not highly hygroscopic  $\text{I}_2\text{O}_5$ . To the contrary, in their chamber work on the reactivity of the  $\text{I}_2/h\nu/\text{O}_3$  system, Saunders and Plane (2006) [25] proposed that polymerization involves  $\text{I}_2\text{O}_5$ , formed by successive oxidation of  $\text{I}_2\text{O}_2$ ,  $\text{I}_2\text{O}_3$  and  $\text{I}_2\text{O}_4$  in the presence of ozone. These authors base their conclusion on an EDX analysis and the calculated average atomic ratio O/I of 2.45 obtained on the sampled iodine oxide particles [25].



Contrary to the work of Saunders and Plane (2006) [25] but in-line with the results presented by Hoffmann et al. (2001) [17] and Jimenez et al. (2003) [18], modeling and theoretical work carried out by Gomez Martin et al. (2007) [23] suggested that the formation of  $I_2O_3$  and  $I_2O_4$  that leads to aerosol formation is best explained by  $\cdot OI$  and  $\cdot OIO$  radical reactivity. In accord with these studies, Saunders et al. (2010) [26] argue that iodine oxide aerosols are formed mainly by polymerization of  $I_2O_4$  with itself and with  $I_2O_3$ . In their experimental flow-tube study of the  $I_2/O_3$  system under "dark" experimental conditions, that is, in the absence of photolyzing radiation, Gomez Martin et al. (2013) argue that  $I_2O_4$  is responsible for the nucleation by polymerization processes, with a minor contribution of  $I_2O_5$  in the early stages of aerosol formation [27]. A hypothesis that is further supported by theoretical studies of  $(I_2O_4)_2$  dimers carried out by Glavez et al. (2013) [28].

There appears to be a convergence of experimental and theoretical results that indicate that the formation of new iodine oxide aerosols results from gas-phase reactions of  $\cdot OIO$  or  $\cdot OI$  radicals to form  $I_2O_3$ ,  $I_2O_4$  or  $I_2O_5$ . It has been presumed that the most stable molecule,  $I_2O_4$  and its dimer, once formed, will undergo polymerization, thus, initiating the aerosol nucleation process [27]. However, a more recent study by Sipilä et al. (2016) suggested that aerosols form *via* sequential addition of  $HIO_3$ , followed by intracuster restructuring to  $I_2O_5$  and recycling of water in the atmosphere or on dehydration [29]. The additional observations of  $HIO_3$  and the singular role of  $H_2O$  raises new questions about different reaction pathways leading to new iodine oxide particle formation. Under humid conditions, the gas-phase  $OIO\cdot + \cdot OH$  reaction or iodine oxides hydration may lead to form  $HIO_3$ . However, the  $OIO\cdot + \cdot OH$  reaction is an unlikely source of  $HIO_3$  because it is limited by the atmospheric concentrations of  $OH$  radicals. Further, the atmospheric  $OH$  radicals may be scavenged by the  $I\cdot + \cdot OH$  reaction in the early stages of the atmospheric oxidation processes. Also, the  $HIO_3$  atmospheric formation from the iodine oxide hydration reaction is known to be endothermic with a high gas-phase reaction energy barrier, therefore, thermodynamically unlikely to be important. [30].

The exact nature and details of the reactive processes and the exact composition of the iodine oxide that polymerizes remain elusive and uncertain [31, 30]. That is, all the intermediate reactions that lead to aerosol formation and the final aerosol composition remain poorly understood and the chemical composition of the iodine oxide aerosol is a subject of much discussion and controversy [30]. This is worrisome for the Nuclear Safety Industry since tests carried out by the Institute of Radio-protection and Nuclear Safety (IRSN, France) show that the bulk of aerosols released within the confinement zone of the nuclear power plant following a major power-plant

87 accident are in the form of iodine oxides. As a result, detailed chemical and physical charac-  
88 terization of iodine oxide aerosols are needed to better understand the chemical and physical  
89 processes responsible for the atmospheric formation, transport and fate of different fission prod-  
90 ucts to better apply and execute appropriate mitigation strategies to protect the public.

91 In this work, we have performed laboratory chamber investigations of  $\text{CH}_3\text{I}$  photolysis in  
92 the presence of ozone under controlled temperature, reactant gas-phase concentration, humidity  
93 and UV-radiation. The impact of relative humidity on the physico-chemical properties of the  
94 formed iodine oxide particles is investigated. A new analytical approach based on on-line gas-  
95 phase and particle formation and growth analysis and off-line transmission electron microscopy  
96 (TEM) characterization is used. Significant differences in the way the particles form and grow  
97 were observed to depend on the given relative humidity experimental conditions, best explained  
98 by particles' morphological changes. The obtained results give a better insight into the singular  
99 hygroscopic behavior of the particles and the chemical processes that lead to the iodine oxide  
100 particle formation, growth and their final chemical composition.

## 101 **2. Experimental approach**

102 The experimental approach described in this work is similar to the one employed in other  
103 emission studies of environmental hazard interest [32]. The experiments involved a series of off-  
104 line and on-line and continuous analyses of both the gas and aerosol phases using state-of-the-art  
105 analytical methods that include Selective Reagent Ionization Time-of-Flight Mass Spectrometry  
106 (SRI-ToF-MS), High Resolution Time-of-Flight Aerosol Mass Spectrometry (HR-ToF-AMS)  
107 and Scanning Mobility Particle Sizer (SMPS) among others. The experimental details relevant  
108 to this work are given below.

109 All experiments were performed within the double-wall VCE1000 (Vötsch Industrietechn-  
110 nik) electro-polished stainless-steel emission test chamber. The internal chamber volume was  
111 916 L. The temperature within the chamber was maintained constant by allowing a temperature-  
112 controlled air to circulate through the external air-jacket housing located between the inner  
113 chamber volume and the outer wall. All experiments were performed using purified compressed  
114 nitrogen carrier gas. The nitrogen flow rate was 2 standard liters per minute (SLPM) and the  
115 corresponding exponential carrier gas exchange rate was 0.13 chamber volume per hour. The  
116 nitrogen gas was produced using the N2 LCMS-1 nitrogen generator (Claind Srl, Italy) based  
117 on air separation by mechanical means by taking filtered atmospheric air, pressurizing it in an  
118 air compressor and then separating it using the pressure swing adsorption technique to deliver a

continuous stream of dry nitrogen. However, a presence of oxygen could not be excluded. The carrier gas throughput dynamics through the chamber were such that there was always a slight over-pressure within the chamber with respect to the atmospheric pressure to prevent outside air from leaking in. The chamber temperature was measured using the platinum resistance temperature detector (Pt100, Class B) and the chamber humidity was measured using a capacitance hygrometer. The experimental chamber temperature was  $T(^{\circ}\text{C}) = 25 \pm 0.4$  and the experimental relative humidity was preconditioned from dry to 57%RH. The listed uncertainty in temperature was an estimated experimental value. In this work, the humidity sensor that was used had a lower limit humidity value of 2%RH and the humidity accuracy was within  $\pm 2\%$  RH. It was considered that the chamber volume was under dry experimental conditions if the relative humidity sensor reading was  $< 2\%$  RH. The relative humidity within the chamber was measured using the Vötsch Industrietechnik chamber relative humidity probe. The probe's humidity range was % 5 to 95 and the humidity deviation in time was  $\% \pm 1$  to  $\pm 3$ .

The UVP Pen-Ray low pressure mercury lamp (Fisher Scientific, France) was used as a photolytic light source. The pen-ray lamp was fitted into a jacketed quartz cylinder to eliminate the  $\lambda = 184.9$  nm Hg atomic line. Water was allowed to circulate throughout the outer jacket of the cylinder to filter out infrared radiation (heat) while a dry nitrogen gas was allowed to flow through the tube fitted with the Hg-lamp to remove oxygen, thus, eliminating its photolysis leading to unwanted ozone formation within the lamp housing. Even though a UV lamp was used as a source of photolyzing radiation, Hughey et al. (2021) [33] reported the  $\text{CH}_3\text{I}$  UV photolysis rate constant to be  $5.92 \times 10^{-4} \text{ s}^{-1}$ , a value that is in agreement with the gamma-irradiation rate, albeit the radiolysis induced  $\text{CH}_3\text{I}$  decomposition kinetics were observed to depend on the initial methyl iodide concentration and the dose rate [34, 33].

As stated above, a state-of-the-art instrumentation park was employed to monitor gas-phase reactant and product temporal profiles and aerosol formation/behavior within the chamber before and after the inception of photolysis. Reactant introduction and the concurrent gas-phase reactant/product concentrations and aerosol measurement experimental details are described in more detail below.

## 2.1. Reactants

The  $\text{CH}_3\text{I}$  (Linde France, Saint-Priest, France) stated minimum purity was 99%. To limit any photo-catalytic or thermal decomposition,  $\text{CH}_3\text{I}$  liquid was stored under dark conditions at  $T = 279\text{K}$ . Deionized water used to humidify the chamber had resistivity greater than  $18 \text{ M}\Omega$ ,

151 TOC < 2 ppb, and was prepared by allowing tap water to pass first through a reverse osmosis  
152 demineralization filter (ATS Groupe Osmose) and then through a commercial deionizer (Milli-  
153 Q, Molsheim, France). The ozone was generated and injected by allowing molecular oxygen  
154 ( $O_2$ , purity > 99.999%, Linde) to flow through a corona discharge generator (CERTIZON C25  
155 Sanders). The oxygen flow rate was  $0.3 \text{ L} \cdot \text{min}^{-1}$  and the resulting ozone flux was  $25 \text{ mg } O_3 \text{ h}^{-1}$ .  
156 In a typical experiment, the desired ozone mixing ratio ( $1.5 \pm 0.1 \text{ ppmV}$ ) within the chamber was  
157 established within ten minutes. This ozone mixing ratio has been chosen so that its concentration  
158 was in excess of that of methyl iodide under the experimental conditions employed. The ozone  
159 mixing ratio was monitored using a commercial ozone analyser (O3 41M; Environment S.A.,  
160 Poissy, France). To correct the flow dynamics requirements of the ozone analyzer ( $2 \text{ L} \cdot \text{min}^{-1}$ ),  
161 an additional and independent flow of dry nitrogen to the analyzer was used in a T-configuration  
162 (2:1 ratio).

163 The gas phase  $CH_3I$  was introduced into the chamber using the gas saturation method first  
164 developed by Regnault in 1845 [35]; one of the oldest and most versatile ways of studying gas-  
165 liquid equilibria involving low vapor pressure compounds. Further, the gas saturation method  
166 used in this work is similar to the one employed in a number of other studies of atmospheric  
167 interest [36, 37, 38, 39] where precise reactant concentrations of volatile or semi-volatile com-  
168 pounds needed to be known. Its operation principle and the calculation of the concentration of  
169  $CH_3I$  leaving the saturation system have been presented in detail elsewhere [40]. Briefly, the  
170 nitrogen carrier gas was allowed to enter the volume containing the  $CH_3I$  liquid that itself was  
171 mixed with glass beads to increase its total surface area and supported on a fritted glass surface.  
172 The saturator volume itself was immersed in a temperature-controlled fluid (water) and kept at  
173 constant temperature ( $T(K) = 279.0 \pm 0.1$ ) and pressure ( $p = 962 \text{ Torr}$ ) using a thermostat with  
174 an accuracy of  $\pm 0.1 \text{ K}$ . The temperature inside the saturator volume was measured using a Type-  
175 J thermocouple (Omega, Manchester, United Kingdom) with an accuracy of  $\pm 0.1 \text{ K}$ . The carrier  
176 gas was allowed to enter the saturator volume, come to a thermal equilibrium with the sample,  
177 and was then allowed to exit through a glass tube. The geometry of the exiting glass tube was  
178 such that the diameter of the glass tube increased with increasing length. This was done to avoid  
179 any sample condensation as the sample and the carrier gas were allowed to leave the saturator  
180 system. The concentration of  $CH_3I$  leaving the saturation system was calculated from the given  
181 vapor pressures [41], mass flow rates, and total pressure within the saturator [40]. Under typical  
182 experimental conditions, the vapor pressure of  $CH_3I$  within the saturator was calculated to be  
183  $183 \pm 4 \text{ Torr}$  and the gas phase  $CH_3I$  mixing ratio within the saturation system was about 19%.

184 This vapor pressure was further diluted with a dilution ratio of 1500:1 using dry nitrogen carrier  
185 gas that was allowed to flow through a system of mass flow controllers before its introduction  
186 into the chamber. Under typical experimental conditions, the reactant introduction lasted about  
187 22 minutes to achieve a homogeneous chamber mixing ratio of  $[\text{CH}_3\text{I}] = 102 \pm 5$  ppbV. This  
188 methyl iodide mixing ratio is on the lower side of what has been modeled to be emitted during  
189 a major nuclear power plant accident [34].

## 190 2.2. *On-line physical and chemical analyses*

### 191 2.2.1. *Gas phase chemical analysis*

192 A commercial Selective Reagent Ionisation Time-of-Flight Mass Spectrometer (SRI-ToF-  
193 MS 8000 Ionicon Analytik GmbH, Innsbruck, Austria) was used to measure the gas-phase  $\text{CH}_3\text{I}$   
194 and other gas-phase products that include selected iodide oxides. Similar to our previous work  
195 on detection of gas-phase methyl iodide, the chemical ionization (CI) reagent was the molecular  
196 oxygen ion ( $\text{O}_2^+$ ) [40]. The drift-tube pressure ( $p_{\text{drift}}$ ) was 2.10 mbar, drift tube voltage was 544  
197 V and the drift tube temperature was held constant at  $T=333.15\text{K}$ . The corresponding  $E/N$  ratio  
198 was 127 Td. ( $1\text{Td}=10^{-17} \text{ V} \cdot \text{cm}^2$ ) where  $E$  is the electric field strength applied to the drift tube  
199 and  $N$  is the buffer gas density. Data were analyzed using the Tofware software V2.5.10.

### 200 2.2.2. *Granulometric assessment*

201 Particle size distribution was measured using the Scanning Mobility Particle Sizer (SMPS).  
202 The SMPS was equipped with a TSI model 3077A 85Kr neutralizer (TSI Inc.), a long col-  
203 umn cylindrical Differential Mobility Analyzer (DMA, TSI 3081) and a Condensation Particle  
204 Counter (CPC, TSI 3776). The DMA sheath air flow was 15 liters per minute (LPM) and the  
205 aerosol flow was 1.5 LPM. The aerosol mobility time resolution within the DMA was 135 s  
206 over  $5.94 < d_m (\text{nm}) < 224.7$ , where  $d_m (\text{nm})$  is the aerosol electric mobility diameter in units of  
207 nanometers, nm ( $1 \text{ nm} = 10^{-9} \text{ m}$ ), and was obtained by changing the voltage exponentially from  
208  $-30 \text{ VDC}$  to  $-7000 \text{ VDC}$ . Similar to the situation listed above for the ozone analyzer, to correct  
209 the flow dynamics requirements of the SMPS sampling flow, an additional and independent flow  
210 of filtered room air (2:1 ratio) was used in a T-configuration.

### 211 2.2.3. *Aerosol physio-chemical analysis*

212 The non-refractory aerosol fraction (evaporates rapidly at  $T=600^\circ\text{C}$  under vacuum condi-  
213 tions) was analyzed *in-situ* and continuously using the HR-ToF-AMS. The AMS analysis al-  
214 lowed to chemically characterize both the non refractory organic and inorganic fraction of

aerosols in the aerodynamic diameter size ranging from  $\sim 70$  nm to 1000 nm. The instrument was used under manufacturer's standard operating conditions (vaporizer temperature was  $600^{\circ}\text{C}$  and the electron impact (IE) detector voltage was 70 eV). The AMS mass spectra analyses of the aerosols allowed to distinguish the iodinated species from the organic fraction. Data were pre-processed using the standard ToF-AMS analysis toolkit SQUIRREL V.1.200. High-resolution peak fitting and integration procedures were performed using the PIKA software (v1.600). Typically, AMS mass spectra are measured using the nitrate equivalent mass temporal mass evolution or nitrate equivalent mass [42]. That is, ammonium nitrate particles are used to calibrate the aerosol mass spectrometer because their physical properties such as density and shape are well established, and they are known to fully vaporize within the ionization zone. Consequently, the AMS mass response to ammonium nitrate signal is used to calculate the "ionization efficiency", that is, number of ions produced per molecule. Therefore, when the ionization efficiency of specific compound is unknown, it is a common practice to express their quantity in "equivalent nitrate" concentrations.

### 2.3. *Off-line analysis*

To evaluate the physio-chemical and morphological nature of the formed aerosols, off-line point samples were collected by filtration using the Mini-Particle Sampler (MPS) [43]. Here, sampled air was allowed to flow through the Transmission Electron Microscope (TEM) TEM-porous grids (Quantifoil, AGAR Scientific, 1.2/1.3 400 mesh). This procedure allowed to sample the aerosols directly on a given support, thus, minimizing any additional sample preparation or sampling artifacts [43]. The filtration flow rate was  $0.3 \text{ L min}^{-1}$  and the sampling efficiency was known. The filtered aerosols were then characterized using the TEM (JEOL JEM2010 microscope, CINaM Electron Microscopy Service, Marseille, France) fitted with a  $\text{LaB}_6$  electron gun under a 200 kV accelerating voltage coupled with energy-dispersive X-ray and EDX spectrometer (BRUKER, Quantax) equipped with a XFlash<sup>®</sup> silicon drift detector.

## 3. Results

The performed experiments are summarized in Table 1. Experiments D0 and D1 were carried out under dark conditions, that is, the UV-lamp was turned off, in the absence and presence of  $\text{CH}_3\text{I}$  at a given  $\text{O}_3$  mixing-ratio value. The Li ( $i = 0$  to 4) and  $\text{Li}^*$  ( $i = 2,3,4$ ) experiments were carried out in the presence of the photolyzing UV-radiation (UV-lamp was tuned on) at given gas-phase  $\text{CH}_3\text{I}$  and  $\text{O}_3$  mixing-ratio values and given relative humidity conditions. A

Table 1: Summary of the experiments performed under "D" dark conditions and "L" in the presence of UV-radiation or "light". T(°C) = 25 ± 0.4. Listed uncertainties are experimental estimates.

Experiments <sup>†</sup>	[CH <sub>3</sub> I] ppbV	[O <sub>3</sub> ] ppmV	%RH	Lamp	Max. Number cm <sup>-3</sup>
D0	0	1.5±0.1	dry, 22±1, 57±3	Off	No particle formation
D1	102±5	1.5±0.1	dry, 22±1, 57±3	Off	No particle formation
L0	0	1.5±0.1	dry, 22±1, 57±3	On	No particle formation
L1	102±5	0	22±1, 57±3	On	Not investigated
L2	102±5	1.5±0.1	dry	On	1.2 × 10 <sup>6</sup>
L2*	not re-injected <sup>‡</sup>	1.5±0.1	dry	On	No particle formation
L3	102±5	1.5±0.1	22±1	On	2.2 × 10 <sup>6</sup>
L3*	not re-injected <sup>‡</sup>	1.5±0.1	22±1	On	1.7 × 10 <sup>4</sup>
L4	102±5	1.5±0.1	57±3	On	1.8 × 10 <sup>6</sup>
L4*	not injected <sup>‡</sup>	1.5±0.1	57±3	On	3.6 × 10 <sup>4</sup>

<sup>†</sup> The L2\*, L3\* and L4\* experiments were performed in series after the experiments L2, L3 and L4, respectively, when the O<sub>3</sub> and CH<sub>3</sub>I mixing ratios have been observed to reach the corresponding instrument lower detection limit values. That is, the L2\*, L3\* and L4\* experiments were carried out without prior chamber cleaning procedure and following the re-injection of O<sub>3</sub>.

<sup>‡</sup> Below the limit of quantification, LOQ < 0.23 ± 0.10 ppbv [40]

cleaning procedure was used prior to every experiment with the exception of Li\* (i = 2,3,4) experiments. Briefly, the internal walls of the chamber were first wiped clean with deionized water using kimwipes. Then, a flow of purified nitrogen gas was allowed to flow through the chamber volume at a rate of 4 liters per minute and was maintained for more than 72 hours. The cleanliness of the chamber was then verified for particle formation using the SMPS and AMS and an organic contamination using the SRI-ToF-MS under given experimental conditions that included a presence and absence of UV-radiation and different relative humidity values. As listed in Table 1, the L0 experiments were carried out in the absence of CH<sub>3</sub>I at a given O<sub>3</sub> mixing-ratio value and L1 experiments were carried out in the absence of O<sub>3</sub> at given CH<sub>3</sub>I mixing-ratio values. The influence of relative humidity in the presence of both gas-phase reagents, namely, CH<sub>3</sub>I and O<sub>3</sub>, were tested in the experiments L2, L3 and L4. As shown in Table 1, experimental conditions that lead to particle formation require the presence of gas-phase methyl iodide, ozone and UV-radiation.

### 259 3.1. $\text{CH}_3\text{I}$ temporal profile

260 To evaluate the gas-phase loss of  $\text{CH}_3\text{I}$  within the chamber, its temporal evolution was stud-  
 261 ied using the SRI-ToF-MS. Some typical normalized  $\text{CH}_3\text{I}$  temporal profiles observed following  
 262 the introduction of  $\text{CH}_3\text{I}/\text{H}_2\text{O}/\text{N}_2$  and  $\text{CH}_3\text{I}/\text{O}_3/\text{H}_2\text{O}/\text{N}_2$  mixtures into the chamber in the pres-  
 263 ence and absence of photolysis are shown in Figure 1.

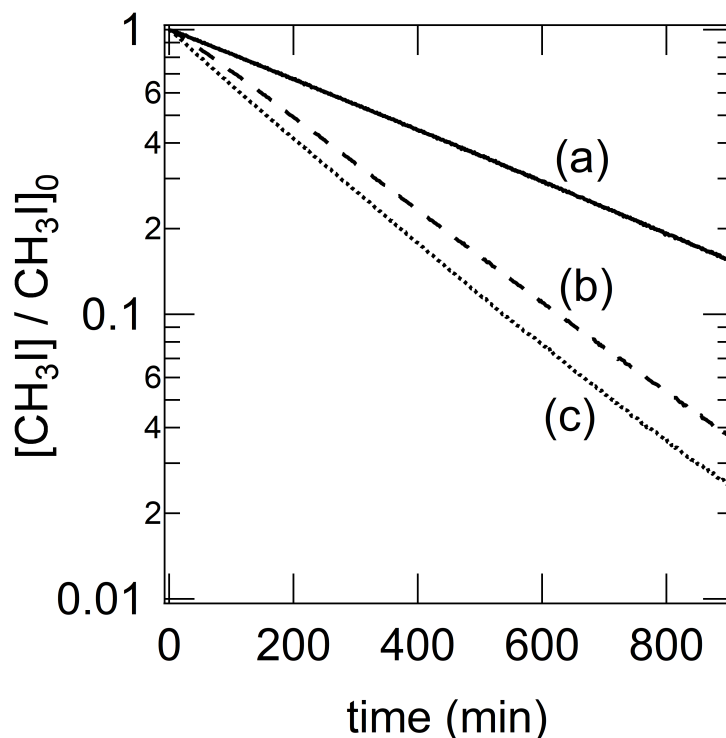


Figure 1: Typical temporal evolution of the concentration of  $\text{CH}_3\text{I}$  observed using SRI-ToF-MS, with the UV-lamp off and on with and without  $\text{O}_3$ . Experimental conditions: (a) Experiments D1, UV-lamp off,  $[\text{CH}_3\text{I}] = 102 \pm 5$  ppb,  $[\text{O}_3] = 1.5 \pm 0.1$  ppm, (b) Experiments L1, UV-lamp on,  $[\text{CH}_3\text{I}] = 102 \pm 5$  ppb, (c) Experiments L2, L3 and L4, UV-lamp on,  $[\text{CH}_3\text{I}] = 102 \pm 5$  ppb,  $[\text{O}_3] = 1.5 \pm 0.1$  ppmV

264 As shown in Figure 1, the  $\text{CH}_3\text{I}$  temporal profiles were observed to be exponential. As a  
 265 result, pseudo-first-order conditions were assumed. The kinetic data shown in Figure 1 pro-  
 266 vide information on the  $\text{CH}_3\text{I}$  gas-phase loss rate within the chamber due to photolysis, wall  
 267 loss or secondary reactions. From the kinetic data using  $\text{CH}_3\text{I}/\text{O}_3/\text{H}_2\text{O}/\text{N}_2$  mixtures in the  
 268 absence of a photolytic light source (Experiments D1), the obtained decay rate constant was  
 269  $k_{\text{obs}} = (3.3 \pm 0.1) \times 10^{-5} \text{ s}^{-1}$ ,  $1\sigma$ . This value corresponds well with the chamber  $\text{N}_2$  renewal flow  
 270 rate of  $2 \text{ L min}^{-1}$  or exponential rate of  $3.6 \times 10^{-5} \text{ s}^{-1}$ . As a result, under the experimental condi-  
 271 tions employed in this work, it was assumed that the gas-phase  $\text{CH}_3\text{I}$  does not react with ozone  
 272 in the absence of photolytic radiation and the wall losses may be considered to be negligible.



From the kinetic data using  $\text{CH}_3\text{I}/\text{H}_2\text{O}/\text{N}_2$  mixtures in the presence of a photolytic light source (Experiments L1) we obtain  $k_{\text{obs}} = (6.3 \pm 0.1) \times 10^{-5} \text{ s}^{-1}$ ,  $1\sigma$ . On the other hand, from the kinetic data using  $\text{CH}_3\text{I}/\text{O}_3/\text{H}_2\text{O}/\text{N}_2$  mixtures in the presence of a photolytic light source (experiments L2, L3 and L4) we obtain  $k_{\text{obs}} = (7.2 \pm 0.4) \times 10^{-5} \text{ s}^{-1}$ ,  $1\sigma$ . From the calculated  $k_{\text{obs}}$  values we conclude that the gas-phase  $\text{CH}_3\text{I}$  is mainly consumed by photolysis. The kinetics of  $\text{CH}_3\text{I}$  loss in the presence of a photolytic light source appears to be independent of humidity but dependent on the ozone mixing ratio within the chamber.

### 3.2. Particle formation and growth

A typical SMPS temporal evolution of the particle number size distribution as well as the corresponding average electric mobility diameter values ( $d_m$ , solid line) are shown in Figure 2. The  $d_m$  was obtained using the aerosol total mass distribution data shown in Figure 3. The UVP lamp was switched on at time  $t_0$  ( $t_0 = 0 \text{ s}$ ), initiating the photooxidation of  $\text{CH}_3\text{I}$  inducing particle formation as shown in Figure 2 and Figure 3.

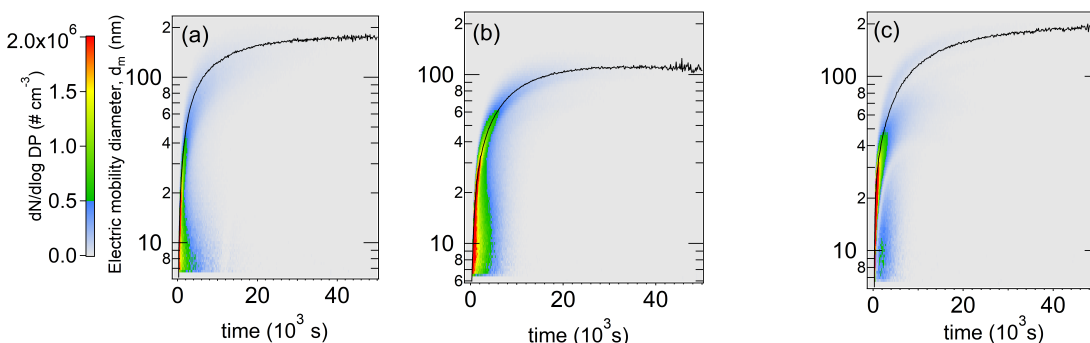


Figure 2: A typical temporal evolution of the total particle number size distribution and the corresponding average electric mobility diameter ( $d_m$ , solid line) obtained using the SMPS. The  $d_m$  was calculated from the mass distribution measurements. The pattern of the matrix illustrates the initiation of particles followed by the particle growth evolution from the nucleation mode to the accumulation mode. Experimental conditions: (a) L2: < 2%RH, (b) L3: RH = 22%RH, (c) L4: 57%RH.

Particle number size distributions were measured between 5.94 and 224.7 nm every 135 seconds. The  $d_m$  were observed to evolve exponentially to reach their maximum size after 7 to 8 hours of reaction time, a range that represents a typical temporal evolution of particle formation observed in the simulation chamber or in the real atmosphere [44]. As shown in Figures 2a and 2b, under dry experimental conditions (L2: < 2%RH) and at a low relative humidity value (L3: 22%RH), the nucleation mode was observed to persist for up to 3 hours with an average aerosol electric mobility diameter of  $\approx 8 \text{ nm}$  and  $14 \text{ nm}$ , respectively. However, as may be seen in

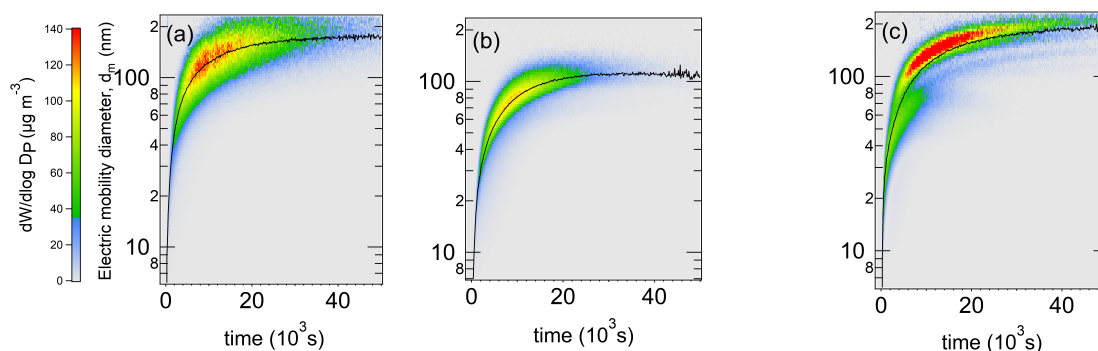


Figure 3: A typical temporal evolution of the total particle mass size distribution and the corresponding average electric mobility diameter ( $d_m$ , solid line) calculated from the mass distribution. The data has been corrected using the density values presented in section 3.4. The shape of the matrix shows the evolution of the accumulation mode. Experimental conditions: (a) L2: < 2%RH, (b) L3: 22%RH, (c) L4: 57%RH.

Figure 2c, at higher relative humidity value (L4: 57%RH), the nucleation contribution to aerosol number quickly evolved to become negligible and gave rise to the accumulation mode of the aerosol growth.

A typical temporal evolution of the total particle mass size distribution and the corresponding  $d_m$  is shown in Figure 3. Here, the shape of the banana-plot illustrates the evolution of the aerosol accumulation mode. Under dry experimental conditions (L2: < 2%RH), the aerosol formation evolve toward a broad particle size distribution regime with an observed maximum average electric mobility diameter of  $173 \pm 6$  nm (Figure 3a). At 22%RH, the aerosols have been observed to evolve toward a narrower particle size distribution with a smaller maximum average mobility diameter of  $103 \pm 8$  nm (Figure 3b). At 57%RH, two accumulation modes were observed to form, evolving toward two distinct maximum average electric mobility diameters of  $\approx 138 \pm 8$  nm and  $205 \pm 8$  nm with the average electric mobility diameter value of  $(192 \pm 8)$  nm (Figure 3c). The two aerosol accumulation modes were observed to distinguish themselves at  $t = 80$  to  $100$  minutes ( $4.8 \times 10^3$  s to  $6 \times 10^3$  s, Figure 3c) when the average electric mobility aerosol diameter exceeded 70 nm.

The nitrate equivalent mass measured using the AMS is shown in Figure 4. As discussed before, ammonium nitrate particles are used to calibrate the aerosol mass spectrometer because their physical properties such as density and shape are well established, and they are known to fully vaporize within the ionization zone [42].

The aerosol fraction is observed to be constituted of iodides, organics and, to a lesser extent, sulfates, nitrates and ammonium ions that are believed to represent residual impurities present within the chamber. The aerosol chemical composition total mass time evolution is observed

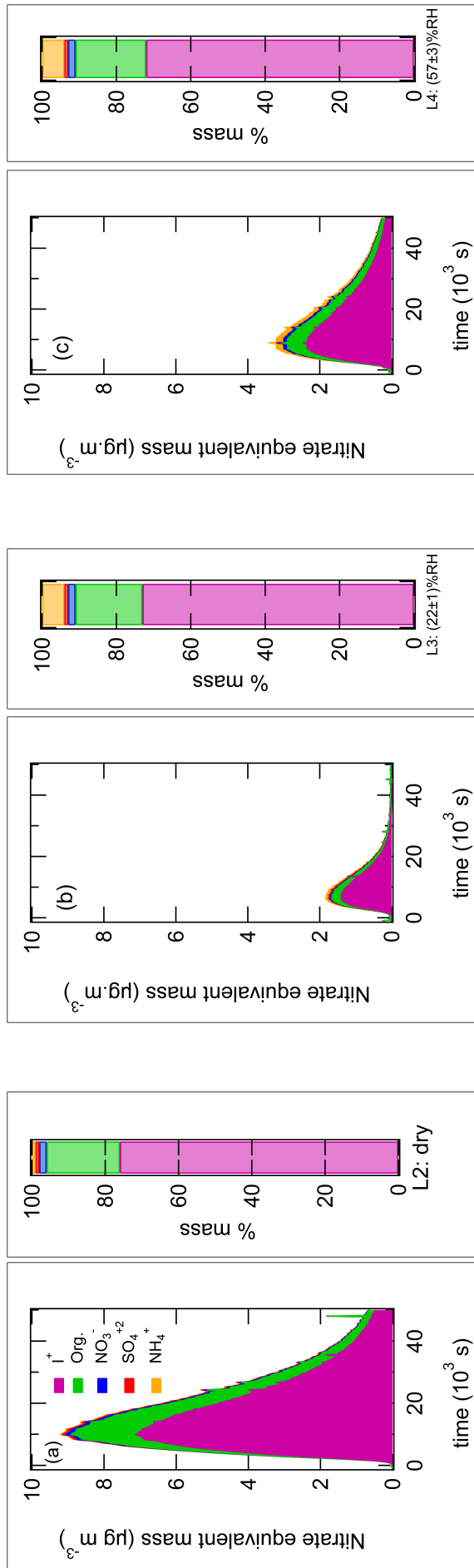


Figure 4: Example of the temporal evolution of the total mass (nitrate mass equivalent) measured by the AMS as well as the proportional evolution ratio to the total mass of each family of compounds within the aerosol. Experimental conditions: (a) L2: < 2%RH, (b) L3: 22%RH, (c) L4: 57%RH.

315 to remain relatively stable as soon as the AMS aerosol detection limit is reached (vacuum aerodynamic mobility diameter,  $d_{va} > 70$  nm). This means that after the aerosols' formation, their  
 316 composition remains homogeneous during growth. As shown in Figure 4, an increase in humidity  
 317 results in the total mass decrease. However, between 22%RH and 57%RH an increase in  
 318 the total mass is observed. This phenomenon is thought to be due to the impact of humidity on  
 319 particle size and morphology. This important result is explained in section 3.5.

321 On the average, the aerosol iodine component fraction represents  $73 \pm 4\%$  of the total mass  
 322 concentration followed by the organic fraction ( $19 \pm 4\%$ ) and traces of ammonium (1 to 6%),  
 323 sulfates and nitrates (1 to 2%). The signal of iodides is dominated by the ion fragment  $I^+$  ( $m/z$   
 324 126.90) followed by  $I_2^+$  ( $m/z$  253.80). In addition to  $I^+$  and  $I_2^+$ , the fragments attributed to the  
 325 family of iodine compounds are the following, in order of importance,  $IO^+$  ( $m/z$  142.89),  $IH^+$   
 326 ( $m/z$  127.91),  $IO_2^+$  ( $m/z$  158.89),  $HIO^+$  ( $m/z$  143.90),  $I_2O^+$  ( $m/z$  269.80),  $I_2O_3^+$  ( $m/z$  301.79),  
 327  $HIO_3^+$  ( $m/z$  175.89),  $HIO_2^+$  ( $m/z$  159.90) and  $I_2O_2^+$  ( $m/z$  285.79).

328 The organic fraction of aerosols may be attributed mostly to the  $CO_2^+$  fragment ( $m/z$  43.98,  
 329 org44) (and by default to the  $CO^+$  fragment ( $m/z$  27.99, org28, according to the standard fragmentation  
 330 approach:  $org28 = org44 \times 1$ )) which represents the most oxidized fraction of organic  
 331 compounds. These two fragments account for  $96 \pm 1\%$  of the total mass attributed to the aerosols' organic  
 332 fraction, independent of the experimental relative humidity conditions. The remaining organic  
 333 fraction mass spectra fragments include  $CH_3^+$  ( $m/z$  15.02),  $CH_2O^+$  ( $m/z$  30.01),  
 334  $C_2H_3O^+$  ( $m/z$  43.01) and  $CHO_2^+$  ( $m/z$  44.99).

335 Similar to the situation described by Pieber and coworkers (2016), it is assumed that the  
 336 reactive behavior of the iodine oxides is equivalent to the  $NH_4NO_3$  and  $(NH_4)_2SO_4$  salt reactivity  
 337 within the AMS ionization source [45]. As a result, it is assumed that the contribution of the  
 338 highly oxidized  $CO_2^+$  and  $CO^+$  fragments to the total aerosol content may be ignored and that  
 339 the aerosols formed under given experimental conditions are principally composed of iodine  
 340 oxides. Further, possible origins of other organic fragments include residual organic compounds  
 341 present within the AMS ionization source that include  $CH_3I$  and its organic radicals formed  
 342 following the photooxidation of  $CH_3I$  and to a much lesser extent the photooxidation of other  
 343 trace organic impurities present in our experimental chamber.

### 344 3.3. Particle morphology

345 Off-line samples were collected to better interpret and understand the aerosols' morphology  
 346 and physico-chemical information that includes particle number and size distribution and par-

347 ticle mass concentration, acquired using the SMPS and AMS techniques. Here, samples were  
348 collected by filtration on porous TEM grids at a specific time under given experimental condi-  
349 tions. A typical TEM image of a sample taken over a 6 minute time interval after 3 hours of  
350 reaction time (starting at  $t \approx 3$  hours) is shown in Figure 5. Additional off-line samples were col-  
351 lected over a period of 6 minutes after 7 hours of reaction time (starting at  $t \approx 7$  hours). At this  
352 moment, the aerosols have been observed to approach their maximum electric mobility diameter.  
353 Typical TEM images observed at three different relative humidity values are shown in Figure 5.  
354 Compared to  $t \approx 3$  hours, the morphology of the collected particle remained unchanged.

355 As shown in Figure 5, the morphology of the collected particles was observed to be explicitly  
356 distinct at different relative humidity values. That is, the morphology of the particles formed at  
357  $< 2\%RH$  and at  $22\%RH$  are observed to be mostly compact or near-spherical in nature. At  $57\%$   
358  $RH$  the formed particles are observed to be mostly like-crystal in nature. Particles collected at  
359  $22\%RH$  (Figure 5) were observed to be clearly smaller in size compared to the particles collected  
360 at  $< 2\%RH$  and at  $57\%RH$  (Figure 5). This result is in accord with the SMPS analysis shown in  
361 Figure 3.

362 As shown in Figure 6 the collected particles may be classified into 4 series. Also, the cor-  
363 responding particle number size distributions obtained using the SMPS averaged over the TEM  
364 grid sampling period (three scan from  $t \approx 7$  hours) is shown in Figure 6 .

365 The Series 1 taken under dry conditions may be attributed to the accumulation mode 1. In a  
366 similar manner, the Series 2 observed at  $22 \pm 1\%RH$  may be attributed to the accumulation mode  
367 2 and the Series 3 observed at  $57 \pm 3\%RH$  may be attributed to the accumulation mode 3 (Figure  
368 6). Since the SMPS measurements operate in the transition regime [46], the elongated particles  
369 observed under dry condition are influenced by their shape and orientation within the Dimen-  
370 sional Mobility Analyzer (DMA) electric flow fields which in turn contribute to the broadening  
371 of the corresponding particle size distributions [47]. Likewise, depending on aerosols' orien-  
372 tation within the DMA electric field flow, a selection of the series 4 aerosols, may be divided  
373 between the accumulation mode 3 and the accumulation mode 4.

374 Based on the TEM analyses, the dry-formed Series 1 aerosols induce diffraction fringes and,  
375 therefore, appear to be agglomerates of fine crystals of compact and elongated nature. Unfortu-  
376 nately, we were not able to better study the fine crystals' diffraction fringes under higher magni-  
377 fication because the the particles' instabilities under the electron beam of the TEM. According  
378 to the SMPS measurements (Figure 2a), these aerosols have likely been formed by nucleation  
379 for  $d_m \leq 20$  nm and then evolved to form larger agglomerates.

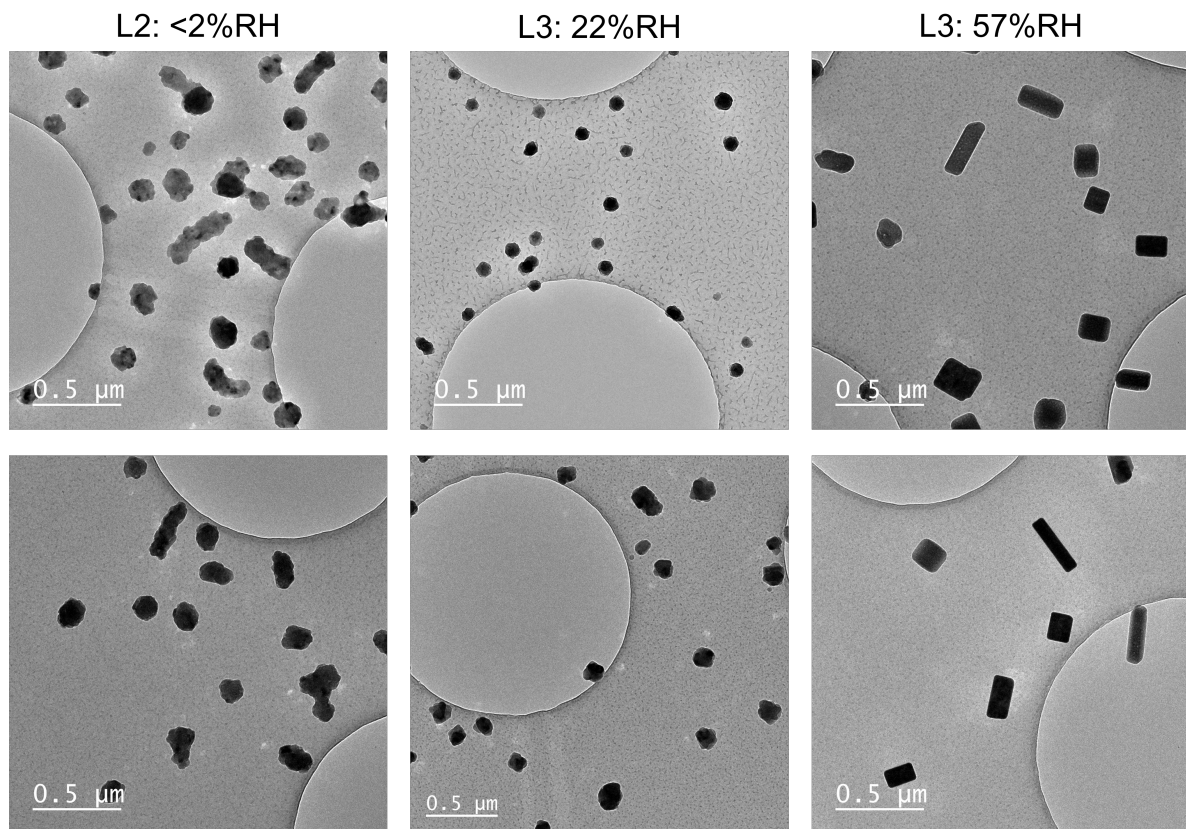


Figure 5: Selected TEM images of particles observed under three relative humidity conditions. Top images: TEM images of samples taken over a period of 6 minutes at  $t \geq 3$  hours reaction time. Bottom images: samples taken over a period of 6 minutes after 7 hours of reaction (from  $t \approx 7$ h).

Series 2 and 3 particles formed at 22%RH and partially at 57%RH have been observed to have more compact structures than those shown in Series 1. These structures appear to be intermediate between those of Series 1 and those of the crystals in Series 4. Based on SMPS measurements (Figure 3c), the Series 4 particles were observed to form when  $d_m > 70$  nm. Sequential samples taken on TEM grids showed that crystals in series 4 progressively evolved from a isomeric or tetragonal shape to orthorhombic structures. It is likely that the formation of these crystals is probably due to a sequential addition in a preferential direction of the molecules that make up the crystal lattice.

To better identify the elemental composition of the aerosols, the Energy Dispersive X-Ray (EDX) analysis was performed. The EDX analysis revealed the presence of oxygen and iodine, a result that is in agreement with the AMS aerosol measurements. However, the calculated O/I ratio is not observed to yield any useful information from one particle to another, resulting from the particles' instability under the electron beam.

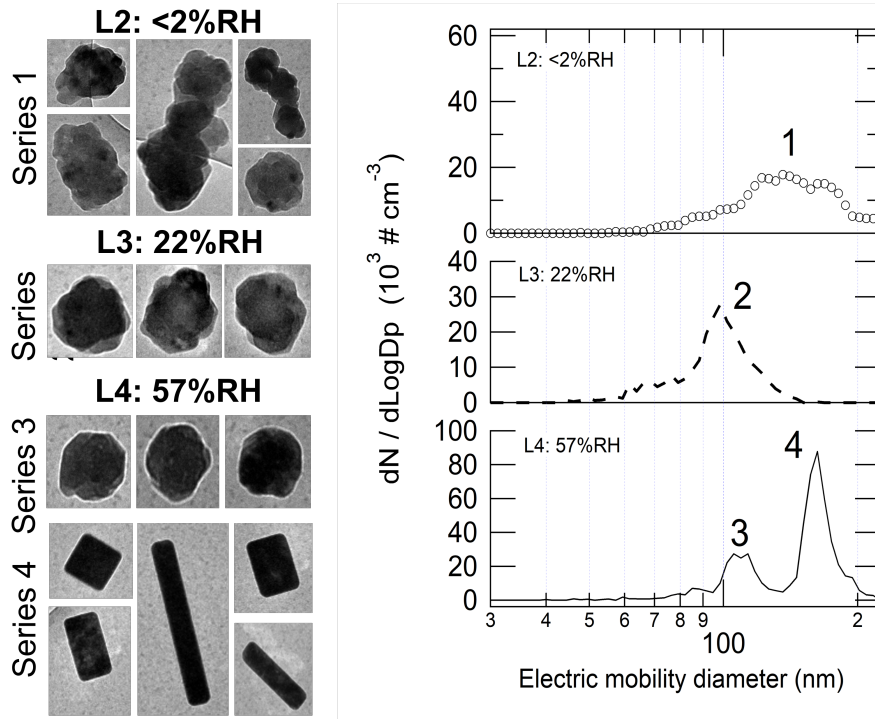


Figure 6: Classification of the observed particles at three different relative humidity values and the SMPS particle number size distributions averaged over the TEM grid sampling period shown in Figure 5. Experimental conditions: (○) open circles, L2: <2%RH; (---) Dashed line, L3: 22%RH; (—) Solid line, L4: 57%RH. The presented particle TEM images do not respect the original scale and are shown only for their morphological structure and shape.

### 3.4. Particle density

To obtain a mass concentration from SMPS measurements, the density  $\rho_p$  of the measured particles must be defined. This information may be obtained from the ratio of the vacuum average aerodynamic diameter measured using the AMS ( $d_{va}$ ) and the average mass-weighted electric mobility diameter measured using the SMPS ( $d_m$ ), equation 1 [18, 46],

$$\rho_p = \rho_0 \frac{d_{va}}{d_m} \quad (1)$$

where  $\rho_0$  is the unit density in  $\text{g/cm}^3$  ( $\rho_0 = 1 \text{ g/cm}^3$ ). The equation 1 is valid only if the particles are spherical or close to spherical. For non-spherical particles, this relation depends on several considerations related to the geometrical shape of these particles. In this case, the density measured with equation 1 is a so-called effective density  $\rho_{\text{eff}}$ , relative to a spherical particle that would have the same electric mobility diameter as well as its aerodynamic equivalent in vacuum [18, 46].

The plot of the average aerodynamic mobility diameter obtained from the dominant fragment  $\text{I}^+$  ( $m/z$  126.90) measured using the AMS as a function of the average mass-weighted electric

mobility diameter,  $d_m$ , measured using the SMPS at three relative humidity values is shown in Figure 7.

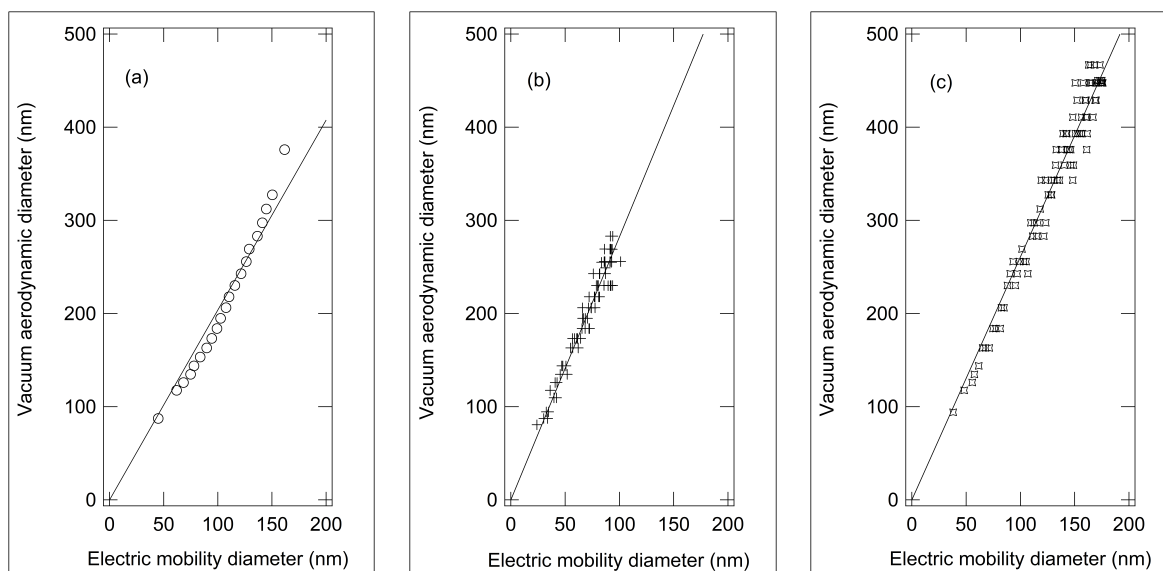


Figure 7: Plot of the average aerodynamic mobility diameter under vacuum obtained from the main fragment  $I^+$  (AMS measurement at  $m/z$  126.90) as a function of the average mass-weighted electric mobility diameter (SMPS measurement). Experimental conditions: (a) L2:  $< 2\%RH$ ,  $\rho_{eff} \approx 2.04 \pm 0.07 \text{ g cm}^{-3}$ , (b) L3:  $22\%RH$ ,  $\rho_{eff} \approx 2.82 \pm 0.02 \text{ g cm}^{-3}$ , (c) L4:  $57\%RH$ ,  $\rho_{eff} \approx 2.61 \pm 0.03 \text{ g cm}^{-3}$ ,  $1\sigma$ . Solid line is the linear correlation plotted from target

Here, a linear correlation is obtained from the diameter data measured using the AMS ( $d_{va} \geq 70 \text{ nm}$ ). The slope of the linear regression line shown in Figure 7 is used to calculate the effective particle density (equation 1).

Based on microscope measurements, particle formed at  $< 2\%RH$  and at  $22\%RH$  are observed to be mostly compact or near-spherical in nature (Figures 5). In these cases, it can be considered that in dry conditions  $\rho_{eff} \approx \rho_p = 2.04 \text{ g}\cdot\text{cm}^{-3}$  and at  $22\%RH$   $\rho_{eff} \approx \rho_p = 2.82 \text{ g}\cdot\text{cm}^{-3}$ .

As shown in Figure 5, at  $57\% RH$  the formed particles are observed to be mostly of crystal structure in nature and the overall effective density  $\rho_{eff} = 2.61 \text{ g}\cdot\text{cm}^{-3}$ .

The effective density values calculated in this work are observed to be lower than the known iodine oxides potentially formed in our system (Table 2) [48, 49, 50]. Otherwise, in their chamber work using the  $\text{CH}_2\text{I}_2/\text{O}_3/h\nu$  reactive system, Jimenez et al. (2003) obtained particle densities of  $0.86 \text{ g cm}^{-3}$  and  $1.22 \text{ g cm}^{-3}$  under dry conditions and values ranging from  $1.6 \text{ g cm}^{-3}$  to  $2.2 \text{ g cm}^{-3}$  at  $65\%RH$ . In the absence of morphological data, these investigators hypothesized that the obtained low density values were a result of the formed particles' aspherical or fractal nature [18]. However, based on the morphological observations reported in our work, we argue that the



Table 2: Physical properties of  $\text{HIO}_3$ ,  $\text{I}_2\text{O}_4$  and  $\text{I}_2\text{O}_5$ .

Name	Formula	Molar mass ( $\text{g mol}^{-1}$ )	Color	Crystal structure	Density ( $\text{g cm}^{-3}$ )	
Iodic acid	$\text{HIO}_3$	175.910	incolor	orthorhombic [51, 52]	4.63	[48]
Diiodine tetroxide	$\text{I}_2\text{O}_4$	317.807	yellow	monoclinic [53]	4.2	[48]
					4.97	[49]
					2.57	[50]
Diiodine pentoxide	$\text{I}_2\text{O}_5$	333.806	white	monoclinic [50]	4.98	[48]

formed particles are globally spherical in nature at  $< 2\%\text{RH}$ ,  $22\%\text{RH}$  and partially at  $57\%\text{RH}$ . Based on our result, the hypothesis of Jimenez et al (2003) only concerns the density values of the elongated particles and crystals but cannot explain the overall observed low density values.

In addition to the morphological considerations, other criteria include particle porosity and particle matrix structure that may contain other materials of lower density. However, the measured particle density was observed to be lower than the density of the materials present within the matrix structure. That is, the AMS analyses concluded that iodine oxides represent by far the majority in the chemical composition of the particles. As a result, we conclude that the particle matrix does not play a significant role in the obtained effective particle density. However, particle porosity may be important, in particular for dry-formed particles because the TEM analysis shows that they appear to be agglomerates of small crystals, favoring a porous structure, resulting in a lower effective density value.

### 3.5. Total mass concentration

The particle total mass concentration was calculated using the SMPS data that includes the particle density  $\rho_p$  and the volume equivalent diameter,  $d_{ve}$  [46].

$$m_p = \frac{\pi}{6} \rho_p d_{ve}^3 \quad (2)$$

Since the particles formed at  $< 2\%\text{RH}$  and  $22\%\text{RH}$  were globally observed to be near-spherical in nature, the volume equivalent diameter ( $d_{ve}$ ) was approximately equivalent to the electric mobility diameter ( $d_m$ ) measured using the SMPS. As a result, it was assumed that the temporal evolution of the total particle mass concentration depended solely on the temporal evolution of the particle number size distribution. At  $57\%\text{RH}$ , the determination of the actual

total particle mass concentration evolution was more challenging due to the crystal shape that directly influence the density and the volume equivalent diameter calculations.

The non-spherical morphology of the collected particles help to explain poor reproducibility of the aerosol total mass measurements from one test to another under given experimental relative humidity conditions (Figure 8). At 57%RH,  $d_{ve}$  is related to  $d_m$  by the dynamic shape factor  $\chi_{t,\Theta}$  which is a function of the flow regime (transition in this case) and the orientation of the particles within the electric flow field of the SMPS [47].

The direct measurement of total mass concentration using the SMPS based on the effective density,  $\rho_{eff} = 2.61 \text{ g/cm}^3$ , is biased. The SMPS total aerosol mass measurement data shown in Figure 8 indicate that the greater the second accumulation mode (Figure 8a) relative to the crystal particle (series 4 crystals presented in Figure 6), the higher the total particle mass (Figure 8b). This phenomenon is known to induce, in some cases, a higher mass reading than the one observed under dry ( $< 2\%RH$ ) conditions (Figure 9a). On the other hand, the corresponding AMS analyses show that the total mass measurement data is always greater under dry conditions than under elevated relative humidity conditions (Figure 9b). Actually, the transition from a dry to humid experimental environment appears to induce a decrease in the particle total mass. Furthermore, the increase in total mass observed between 22%RH and 57%RH is due to the formation of isomeric tetragonal and orthorombic crystals.

The masses measured using the SMPS at 57%RH appear to be overestimated due to the effect of particles' shape on the measured electric mobility diameter within the DMA of the SMPS. This effect is especially important for the more elongated crystal shapes as they travel across the DMA.

On the other hand, the total mass measurements using the AMS are likely to be underestimated. For example under dry conditions, a significant difference between the SMPS and AMS total mass concentration data is observed as shown in Figure 9. AMS data are given in terms of nitrate equivalent mass. A calibration specific to the iodine oxides formed therefore needs to be carried out. To date, it is difficult to establish with certainty the true mass of iodine oxide particles formed at 57% relative humidity. Therefore, we propose that relative humidity inhibits the formation of iodine oxides. This proposition is in-line with the results on the inhibition of iodine oxide formation by relative humidity reported by Saunders et al. (2010) [26] and Galvez et al. (2013) [28]. Further, our results show that at a given value of relative humidity, a well-defined geometric crystal shape formation is favored.

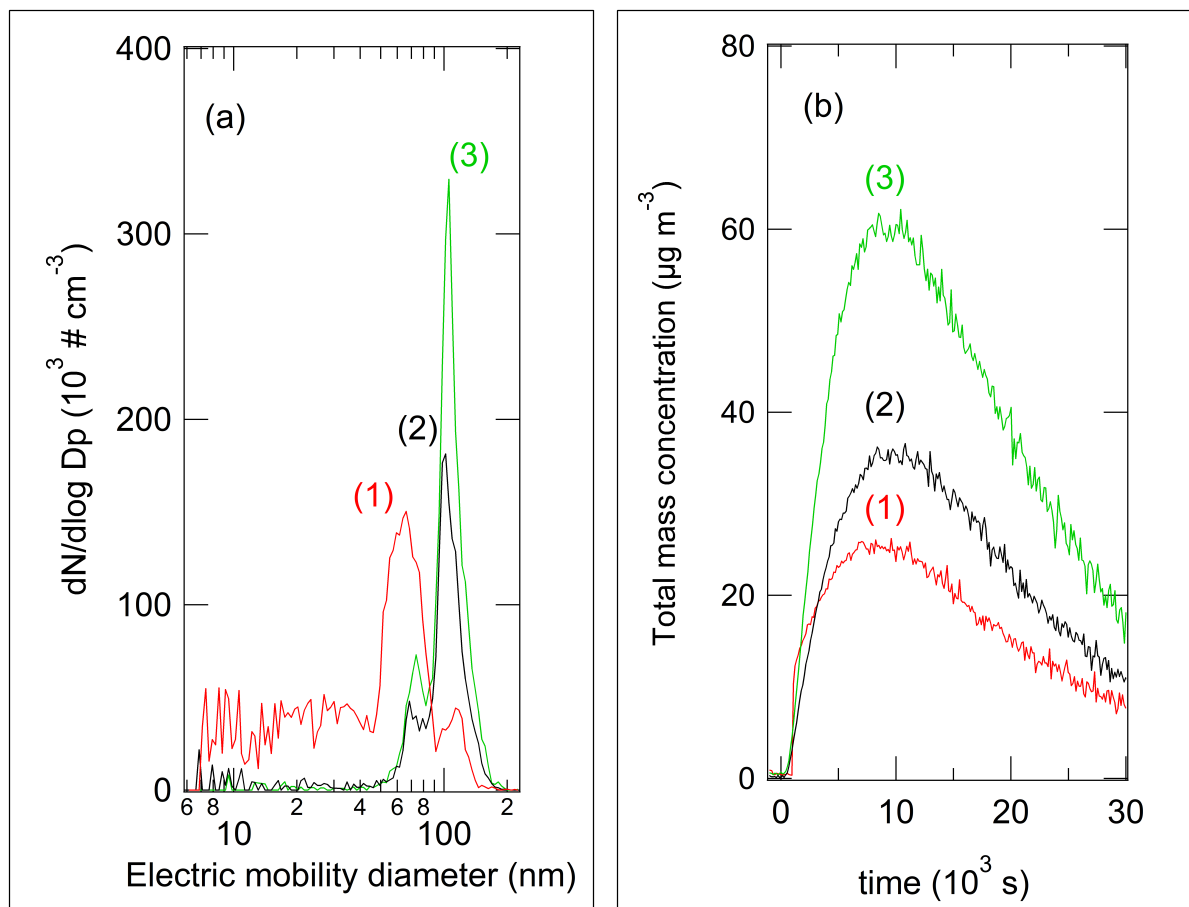


Figure 8: SMPS measurements. Experimental conditions: L4, 57%RH. (a) the particle size distributions in number of particles measured at the maximum of the total particles mass obtained ( $t_{\text{max}}$ ) for 3 tests and (b) The corresponding temporal evolution of the total particle mass concentration.

### 3.6. Particle hygroscopic properties

In their chamber work, Jiménez et al. 2003 [18] investigated the hygroscopic properties of iodine oxide aerosol formation using the  $\text{CH}_2\text{I}_2/h\nu/\text{O}_3$  reactive system. A reduction in the diameter of particles previously formed in a dry atmosphere was observed when they are exposed to moisture. In the case of pre-formed particles at 65%RH, exposure to moisture did not show a significant influence [18]. The authors deduce the formation of  $\text{I}_2\text{O}_4$ , judged to be non-hygroscopic, and explain the shrinkage of the dry-formed particles by the fractal nature of the particles, which collapse by capillary forces induced by water vapor.

To evaluate the hygroscopic behavior of particles formed under given experimental conditions, we performed additional tests on the collected iodine oxide aerosols. Here, the TEM grid samples were collected under given three experimental relative humidity conditions. More, two successive samples were collected under each experimental condition. Then, the first grids were immediately placed in a sealed pocket under dry nitrogen. The others were exposed to a flow of

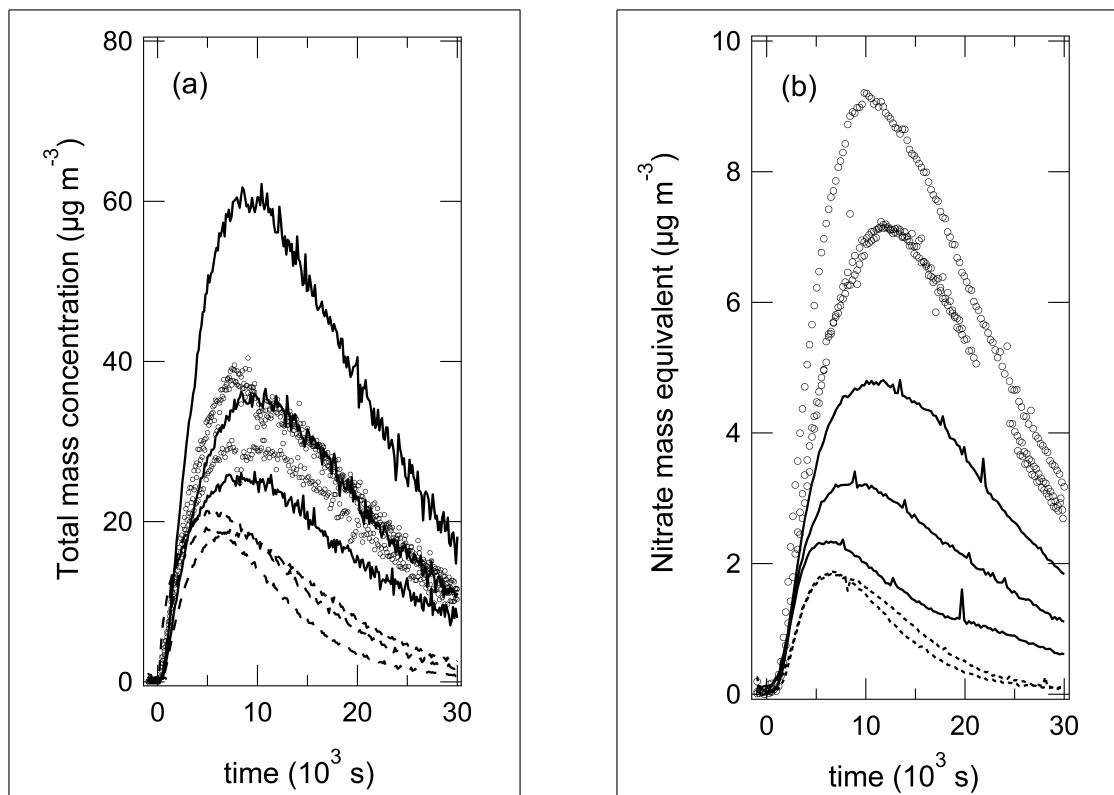


Figure 9: Temporal evolution of the particle total mass concentration under given experimental conditions of relative humidity (a) measured by the SMPS and (b) by the AMS. ( $\circ$ ) open circles, L2: dry ( $<2\%$  RH); (---) Dashed line, L3: 22% RH; (—) Solid line, L4: 57% RH. Separate data curves for each given experimental conditions of relative humidity shown in the two figures are used to better illustrate variability in temporal evolution.

nitrogen gas humidified at  $(83 \pm 1)\%$  RH for 20 minutes.

The particles formed at 22% RH and 57% RH did not show a significant change by exposure to humidity (comparison between the unexposed TEM grid and the one exposed at  $83 \pm 1\%$  RH), thus in agreement with the observations of Jiménez et al. 2003. On the other hand, on dry-formed particles ( $< 2\%$  RH), a clear change is observable as shown in figure 10.

Knowing the TEM grid collection efficiency [43] and the total volume sampled (6 minutes at  $300 \text{ cm}^3 \text{ min}$ ), the particle size distribution may be obtained from TEM images using "ImageJ" image processing and analysis software. Figure 11 shows the particles size distribution obtained from images of the unexposed grid (figure 10a) and distribution form that of exposed to 83% RH (figure 10b) as well as the corresponding SMPS particle size distributions averaged over the sampling time. The diameter used to define the distributions from TEM images is an area equivalent diameter ( $d_{ae}$ ) calculated from the air occupied by a particle. Distributions are obtained between  $d_{ae} = 46 \text{ nm}$  and  $d_{ae} = 368 \text{ nm}$ . The particles of smaller diameters obtained by SMPS were perceptible under TEM on the unexposed grids, however they are very unstable

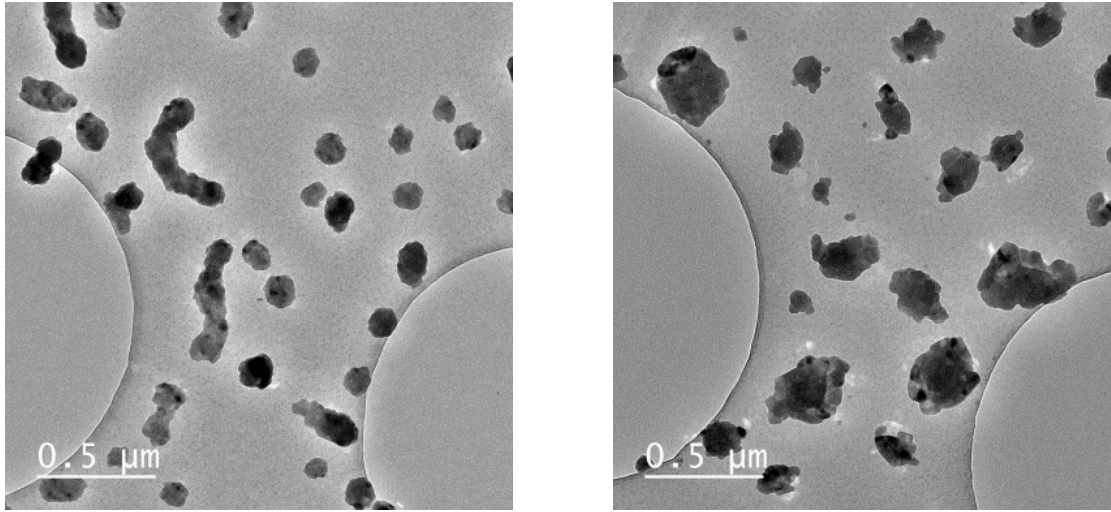


Figure 10: Comparative TEM images of dry-formed and dry-conditioned particles ( $RH < 2\%$ ) with those exposed to  $83 \pm 1\%RH$  for more 20 minutes. The collected particles shown on the left have not been exposed to humidity while the particles shown on the right have been exposed to  $83\%RH$ . TEM sampling time: 6 minutes

502 and disappear rapidly by sublimation before obtaining an image.

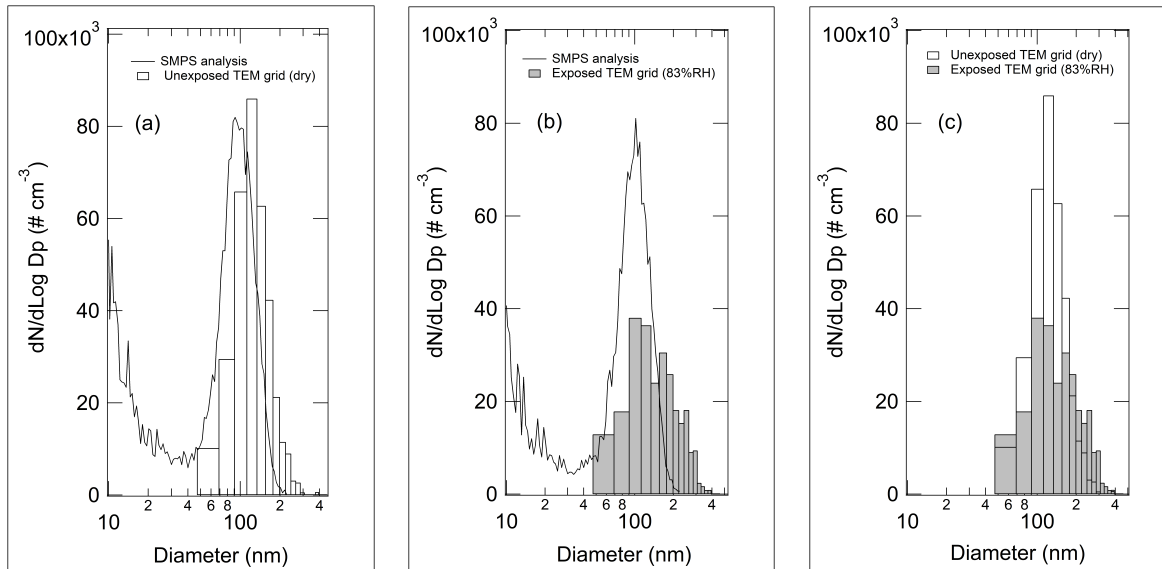


Figure 11: (a) Typical particles size distribution obtained from images of the unexposed grid. (b) Particles size distribution obtained from the TEM grid exposed to  $83\%RH$ . The figures (a) and (b) also show the corresponding SMPS particle size distributions averaged over the sampling time. (c) Comparison of the two particle size distributions obtained from TEM images.

503 As presented in figure 11a, the particle size distribution obtained from the unexposed TEM  
 504 grid is in agreement with that obtained using the SMPS. On the other hand, as shown in figure  
 505 11b, the particle size distribution obtained from the TEM grid exposed to  $83\%RH$  differs. By  
 506 comparing the two distributions (Figure 11c), it appears that exposing the particles collected on  
 507 TEM grids to  $83\%RH$  causes a decrease in the number of particles with a diameter between  $d_{ae}$

508 = 68 nm and  $d_{ae} = 154$  nm and an increase in particles of higher  $d_{ae}$ . As can be seen visually in  
 509 figure 10, the particle morphology is observed to evolve as the TEM grid is exposed to humidity.  
 510 By exposure to  $83 \pm 1\%$  RH, particles of initially elongated morphology probably collapse and  
 511 tend towards a more compact form. It is also probable that rearrangements of particles at the sur-  
 512 face of the TEM grid take place by deliquescence, coagulation and further efflorescence which is  
 513 the origin of the new particle size distribution and morphology observed after exposure to 83%  
 514 RH. Obviously, these observations require further study. However, our observations of collected  
 515 particles on TEM grids reflect the hygroscopic nature of the iodine oxides particles formed under  
 516 dry conditions and particles' non-hygroscopic nature formed under humid conditions.

### 517 3.7. Formation of gas-phase $I_2$

518 Elucidation of a reaction mechanism that leads to the formation of gas-phase molecular io-  
 519 dine is important since such mechanism will give crucial information on reactant precursors,  
 520 namely  $IO\cdot$ ,  $OIO\cdot$ ,  $I_2O_4$  and  $I_2O_5$ , that are involved in aerosol formation. As shown in Table  
 521 1, L0 and Li\* (i= 2,3,4) tests were performed without the  $CH_3I$  reactant under the same ex-  
 522 perimental conditions of ozone mixing ratio, relative humidity, and the presence of photolyzing  
 523 radiation. The difference is that the Li\* tests were carried out in series after the Li tests with-  
 524 out prior chamber cleaning procedure and following the re-injection of  $O_3$ . The L0 and L2\*  
 525 experiments did not result in any particle formation.

526 On the other hand, particle formation was observed in the L3\* and L4\* experiments that  
 527 were carried out at relative humidity values of 22%RH and 57%RH, respectively (Table 1). The  
 528 observed particle formation in the L3\* and L4\* experiments is best explained by the secondary  
 529 formation of gas-phase  $I_2$  which takes place in a humid environment. Indeed, during the L3 and  
 530 L4 tests, the observed formation and evolution of particles at 22%RH and 57%RH, respectively,  
 531 was observed to be accompanied by a release of molecular iodine,  $I_2$ . As shown in Figure 11a,  
 532 the gas-phase SRI-ToF-MS analyses carried out during the L4 experiments at 57%RH revealed  
 533 an important molecular iodine signal. The formed  $I_2$  probably partitions between the gas and  
 534 particle phases and is deposited on the simulation chamber walls. During L3\* and L4\* tests, the  
 535 re-injection of  $O_3$  (arrow 1 in Figure 12b) under UV-visible irradiation induced a release of  $I_2$   
 536 that has been adsorbed on chamber walls (arrow 1 in Figure 12). This process is accompanied  
 537 by the formation and evolution of a small number of particles.

538 Consequently, particles formed during the L3\* and L4\* tests under given experimental con-  
 539 ditions of relative humidity are likely the result of a photo-oxidation of desorbed or volatilized

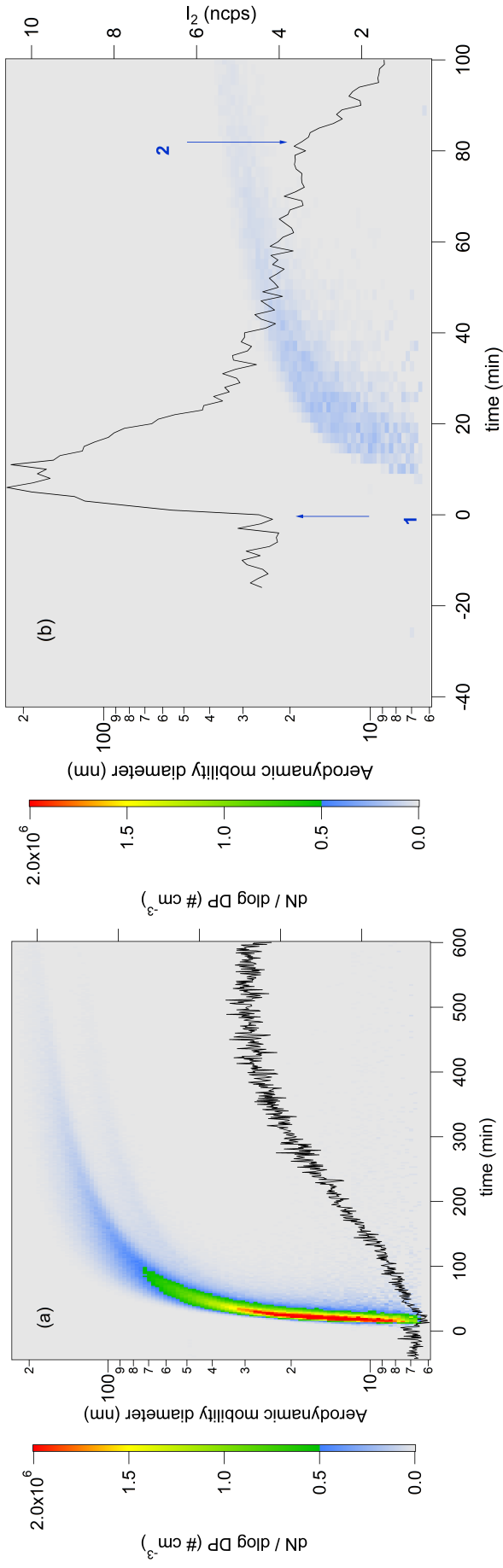


Figure 12: (a) Typical temporal evolution profile of the particle size distribution in number of particles formed according to the L4 tests and evolution of the emissions of molecular iodine  $I_2$  observable by SRI-ToF-MS ( $\%RH = 57 \pm 3$ ). The  $I_2$  signal (solide line) is normalized to  $1 \times 10^6$  cps with respect to the  $O_2^+$  chemical ionization reagent signal (SRI-ToF-MS). (b) Arrow 1 denotes the initiation of the L4\* test where ( $O_3/h\nu$ ) is re-introduced into the chamber, and arrow 2 denotes an increase in the carrier gas renewal rate ( $N_2/O_2$ ) within the chamber for rapid evacuation.

secondary  $I_2$  that has been formed during the L3 and L4 tests and adsorbed by the walls via dry and humid deposition on the chamber walls.

## 4. Discussion

The results obtained in this chamber work using the  $CH_3I/h\nu/O_3/H_2O$  gas-phase system show a formation of new particles. That is, when a known quantity of gas-phase  $CH_3I$  and  $O_3$  was allowed to enter the simulation chamber at a given relative humidity value, aerosols were observed to form as soon as the UV lamp was turned on, inducing a nucleation process. This nucleation process was assumed to be initiated by a direct photolysis of gas-phase  $CH_3I$  and the resulting major fraction of the formed aerosols is identified to contain iodine oxides ( $H_xO_yI_z$ ). The identified  $H_xO_yI_z$  aerosol fraction was observed to form in the aerosols' nucleation mode initiated by the UV radiation and evolve towards the accumulation mode, when the total aerosol mass and number were observed to increase.

The formation of iodine oxide aerosols was studied at three experimental conditions of relative humidity, namely,  $< 2\%RH$  or "dry" conditions,  $(22 \pm 1)\%RH$  and  $(57 \pm 3)\%RH$ , and at atmospheric pressure and ambient temperature. Relative humidity was observed to play a major role in the way particles initially form and evolve in time. An increase in the experimental relative humidity was observed to influence the size and morphology of the formed aerosols and induce a decrease in the total aerosol mass. The aerosols formed in "dry" conditions and at  $22\%RH$  were observed to be compact in nature. On the other hand, the aerosols formation at the experimental relative humidity value of  $57\%RH$  was observed to favor additional crystal formation.

The aerosol chemical composition total mass time evolution was observed to remain relatively stable as soon as the AMS aerosol detection limit was reached ( $d_{va} > 70\text{ nm}$ ). That is, once formed, the aerosol composition remained homogeneous during growth.

The values of the average mass-weighted electric mobility diameter ( $d_m$ ) and the free-molecular regime aerodynamic diameter ( $d_{va}$ ) were used to calculate the effective particle density, a value that is equivalent to the density of spherical particles. This is the case for particles that were observed to form in this work in "dry" conditions ( $< 2\%RH$ ) and at  $22\%RH$ . The calculated particle densities were found to be lower than the particle densities of other known iodine oxides that may have potentially formed in our experimental system [48, 49]. At this point we cannot explain this discrepancy. However, it is well known that the particle density is lower than the density of matter if the volume occupied by the particle is greater than the volume occupied

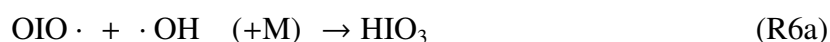


572 by the matter. This is particularly the case for porous agglomerates and particles containing  
 573 vacuum or gases [46].

574 In this work, the off-line TEM analyses of the collected particles indicated a presence of  
 575 diffraction fringes. Unfortunately, we were not able to further explore the observed diffraction  
 576 fringes because of the instabilities of the collected iodide oxide particles under the electron  
 577 beam. Nonetheless, the observed diffraction fringes are indicative of a crystalline and/or semi-  
 578 crystalline character of the collected particles. That is, in dry conditions (< 2%RH), initial  
 579 aerosol nucleation likely results in a production of fine crystals. Then, these fine crystals  
 580 grow to reach an electric mobility diameter of  $d_m \approx 20$  nm and finally agglomerate to reach  
 581 the  $d_m \approx 173$  nm. Under higher relative humidity experimental conditions, namely, 22%RH, the  
 582 aerosols appear to be in an intermediate arrangement state between the near-spherical geometric  
 583 shape and the crystals geometric shape. The collected crystals at 57%RH have been clearly ob-  
 584 served to be of isomeric, tetragonal and orthorhombic geometric shape. These geometric shape  
 585 suggests that the aerosols were formed by sequential addition in a preferential direction of the  
 586 constituent compound. The iodic acid,  $\text{HIO}_3$ , is the likely candidate for the particle's observed  
 587 morphology. That is, it is known that  $\text{HIO}_3$  crystallizes in an orthorhombic space group [51, 52].  
 588 Other iodine oxides ( $\text{I}_2\text{O}_4$  and  $\text{I}_2\text{O}_5$ ) are unlikely candidates for the observed morphology since  
 589 they are known to crystallize in a monoclinic space group that likely favorize needle shape single  
 590 crystals [53, 50].

591 The obtained result in this work is in accord with field studies that reported iodate ions ( $\text{IO}_3^-$ )  
 592 to constitute a major fraction of iodides found within coastal and marine aerosols [54]. Pechtl et  
 593 al. (2007) reported that the main source of  $\text{IO}_3^-$  ions in marine aerosols is gas-phase  $\text{HIO}_3$  and  
 594  $\text{HI}$  [54]. This observation is important since a more recent study by Sipilä et al. (2016) suggested  
 595 that aerosols form *via* sequential addition of  $\text{HIO}_3$ , followed by intracluster restructuring to  $\text{I}_2\text{O}_5$   
 596 and recycling of water through the atmosphere or dehydration [29].

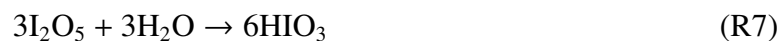
597 It is known that atmospheric gas-phase  $\text{HIO}_3$  may be formed in the reaction of OH radicals  
 598 with OIO radicals and this reaction may occur by several reactive pathways.





599 However, only the reaction pathway R6a that leads to form  $\text{HIO}_3$  is important since the  
600 energy barriers of reactions R6b and R6c are considered to be too high to be of atmospheric  
601 significance [55, 56].

602 While the reaction R6a represents a possible gas-phase source of the observed aerosol  $\text{HIO}_3$   
603 fraction, this reaction is of importance only under higher OH radical concentrations. On the  
604 other hand, the hydrolysis of  $\text{I}_2\text{O}_4$  and  $\text{I}_2\text{O}_5$  is also known to form  $\text{HIO}_3$ . The reaction of  $\text{I}_2\text{O}_5$   
605 with  $\text{H}_2\text{O}$  results in the formation of  $\text{HIO}_3$  (reaction R7) [53, 57] and the hydrolysis of  $\text{I}_2\text{O}_4$  may  
606 forms molecular iodine,  $\text{I}_2$ , and  $\text{HIO}_3$  (reaction R8) [49], .



607 In their work on photochemically generated nanoparticles from gas-phase  $\text{I}_2$  in the presence  
608 of  $\text{O}_3$ , Saunders and Plane (2006) [25] argue that  $\text{I}_2\text{O}_5$  aerosols are formed following a series  
609 of exothermic reactions of  $\text{I}_2\text{O}_2$ ,  $\text{I}_2\text{O}_3$  and  $\text{I}_2\text{O}_4$  with ozone. However, based on theoretical  
610 calculations carried out by Kumar et al. (2018) and Khanniche et al. (2016), the  $\text{I}_2\text{O}_5$  hydration  
611 reaction (R7) is unlikely to occur in the gas phase [31, 57].

612 Experimental and theoretical study carried out by Saunders et al. (2010) [26] and theoretical  
613 study carried out by Galvez et al. (2013) [28] support a hypothesis that  $\text{I}_2\text{O}_4$  is a key compound  
614 involved in the nucleation and formation process of iodine oxides through polymerization of  
615  $\text{I}_2\text{O}_4$  with itself or with  $\text{I}_2\text{O}_y$  ( $y = 3$  to  $5$ ). It has also been shown that the presence of  $\text{H}_2\text{O}$   
616 inhibits the particle formation and growth, a result that is consistent with our observations. A  
617 more recent theoretical study by Martin et al. 2020 [30] suggested that the formation of  $\text{HIO}_3$   
618 particles does not proceed via the slow gas-phase reaction of water with iodine oxides but rather  
619 the particle formation is initiated by progressive hydration clustering of iodine oxides  $\text{I}_x\text{O}_y$ .  
620 These investigators proposed a mechanism that leads to the formation of  $\text{HIO}_3$  by hydration of  
621  $\text{I}_2\text{O}_5$  units present in  $\text{OIO}(\text{I}_2\text{O}_5)_n$  clusters according to the reaction R7. However, the results  
622 obtained in this work that are based on the SRI-ToF-MS analysis suggests a formation of gas-  
623 phase  $\text{I}_2$  in humid conditions whereas  $\text{I}_2$  is not supposed to form during the experiments carried  
624 out in dry conditions. As a result, we argue in this work that  $\text{HIO}_3$  is likely accompanied by a

625 release of molecular iodine  $I_2$ . Therefore, we suspect an additional role of  $I_2O_4$  units according  
626 to the reaction R8 or another iodine oxide cluster that may react with  $H_2O$  to release  $I_2$ .

627 Also, the observed hygroscopic behavior of the particles formed under dry and humid con-  
628 ditions is important. Only dry-formed particles show a noticeable growth and change in size  
629 and morphology under the effect of relative humidity. This finding is consistent with the hygro-  
630 scopic behavior observations of Jimenez et al. (2003) [18]. However, these authors argue that  
631  $I_2O_4$  is central to particle formation under dry and humid conditions. In our work, we argue that  
632  $HIO_3$  is a likely candidate involved in iodine oxide particle formation under humid conditions.  
633 According to Selte et al. (1968) [53], the formation of  $HIO_3$  is irreversible and its stable form  $\alpha$ -  
634  $HIO_3$  can be stored at room temperature and humidity. According to Murray et al. (2012),  $HIO_3$   
635 has a very low hygroscopicity and its deliquescence point is high at %RH = 82.5 [58]. Thus,  
636 particles formed under a dry atmosphere (< 2%RH ) are thought to be composed of porous  
637 agglomerates of hygroscopic fine crystals which, on prolonged exposure at 83%RH, evolve to  
638 form more compact particles of  $HIO_3$ . The same hydration process may occur in the aerosol  
639 phase. As proposed by Martin and co-workers (2020) [30] the nucleation process is thought to  
640 be driven by  $I_xO_y$  cluster reactions. Under dry experimental conditions, the nucleating particles  
641 evolve progressively as agglomerates of small or fine crystals. Under humid atmosphere condi-  
642 tions, water will progressively hydrate  $I_xO_y$  clusters and promote a well-defined  $HIO_3$  geometric  
643 shape crystals of low hygroscopicity.

644 Further, in our chamber experiments we observe that humidity tends to inhibit particle for-  
645 mation and growth. This observation is consistent with the results of Saunders et al. (2010) [26]  
646 and Galvez et al. (2013) [28] who argue that iodine oxides form relatively stable complexes  
647 with water and may block nucleation processes [26, 28]. Our observational results complement  
648 the results obtained by these investigators. That is, the lower total mass that was observed under  
649 humid conditions may also originate from the  $I_2$  loss (reaction R8).

## 650 5. Conclusions

651 Iodine oxide aerosol formation and growth from the photo-oxidation of gas-phase  $CH_3I$  was  
652 studied under controlled laboratory conditions of temperature, relative humidity, reactant gas-  
653 phase concentration and UV radiation. The work aimed to characterize the impact of relative  
654 humidity in the physico-chemical properties of the formed particles and to identify the chem-  
655 ical processes that lead to aerosol initiation and growth. The  $CH_3I$  degradation kinetics were  
656 observed to increase with increasing ozone concentration and did not appear to be significantly

657 influenced by humidity. The chemical analyses of the formed aerosols showed that the particles  
658 were mainly composed of iodine oxides. The temporal evolution of the formed iodine oxide  
659 particle size distribution was observed to change depending on the experimental relative hu-  
660 midity within the chamber. An increase in the experimental relative humidity was observed to  
661 influence the size, morphology and hygroscopicity of the formed particles and induce a decrease  
662 in the total particle mass. Off-line transmission electron microscopy analysis showed that hu-  
663 midity favored the formation of isomeric, tetragonal and orthorhombic crystals whose chemical  
664 composition was likely iodic acid,  $\text{HIO}_3$ , known to crystallize in an orthorhombic space group.

665 The formation of new iodine oxide particles is initiated by the gas-phase recombination or  
666 cross-reactions of  $\text{OIO} \cdot$  or  $\text{IO} \cdot$  radicals to form  $\text{I}_2\text{O}_3$ ,  $\text{I}_2\text{O}_4$  or  $\text{I}_2\text{O}_5$ . These gas-phase recom-  
667 bination or cross-reactions lead to form larger iodine oxide  $\text{OIO}(\text{I}_2\text{O}_5)_n$  clusters. In this work it  
668 was observed that the particles formed under humid conditions are accompanied by the release  
669 of  $\text{I}_2$  which may indicate a potential role of  $\text{I}_2\text{O}_4$  in the particle evolution processes. Under  
670 given experimental conditions, a dry environment was observed to favor the formation of ag-  
671 glomerates of fines crystals. On the other hand, humid conditions were observed to promote the  
672 formation of stable non-hygroscopic crystals of  $\text{HIO}_3$  and the release of  $\text{I}_2$ .

673 The experimental conditions of temperature and relative humidity may be somewhat dif-  
674 ferent from the actual nuclear power plant accident scenario. In view of the results obtained  
675 in this work, it is probable that under given conditions, the iodine oxide particles formed from  
676 iodomethane are of the  $\text{HIO}_3$  crystal type, stable, but likely to deliquesce relative humidity values  
677 above 83%RH. This will likely be accompanied by a secondary formation of gaseous molecular  
678 iodine, itself strongly involved in the formation of new iodine oxide particles.

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